

This version of the article has been accepted for publication, after peer review (when applicable) and is subject to Springer Nature's AM terms of use, but is not the Version of Record and does not reflect post-acceptance improvements, or any corrections. The Version of Record is available online at: <https://doi.org/10.1007/s00216-015-8859-z>
Postprint of: Marć M., Namieśnik J., Zabiegała B., The home-made in situ passive flux sampler for the measurement of monoterpene emission flux: preliminary studies, ANALYTICAL AND BIOANALYTICAL CHEMISTRY, Vol. 407 (2015), pp. 6879–6884

The home-made in-situ passive flux sampler for the measurement of monoterpenes emission flux: preliminary studies

Journal:	<i>Analytical and Bioanalytical Chemistry</i>
Manuscript ID:	ABC-00605-2015.R1
Type of Paper:	Note
Date Submitted by the Author:	n/a
Complete List of Authors:	Marć, Mariusz; Gdańsk University of Technology, Faculty of Chemistry, Department of Analytical Chemistry Namiesnik, Jacek; Gdansk University of Technology, Faculty of Chemistry Zabiegała, Bożena; Gdansk University of Technology, Department of Analytical Chemistry, Chemical Faculty
Keywords:	Air / Gases, Sampling, GC, Organic compounds / Trace organic compounds, Thermal methods

SCHOLARONE™
Manuscripts

Review

1
2
3 1 **The home-made in-situ passive flux sampler for the measurement of monoterpenes**
4 **emission flux: preliminary studies**
5
6
7 3

8 4 Mariusz Marć¹, Jacek Namieśnik¹, Bożena Zabiegała^{1*}
9
10 5

11 6 ¹*Department of Analytical Chemistry, Faculty of Chemistry, Gdańsk University of*
12 *Technology, Gdańsk, Poland*
13
14 7
15 8

16 9 **Address for correspondence: Bożena Zabiegała, Department of Analytical Chemistry,*
17 *Gdańsk University of Technology, ul. Narutowicza 11/12, PL 80-233, Gdansk, Poland.*
18 *e-mail: bozena.zabiegala@pg.gda.pl*
19
20
21 12

22
23 **Abstract**
24
25 14

26 15 The paper presents the construction and metrological characteristics of the home-made in-situ
27 passive flux sampler, an analytical tool representing small-scale emission chambers working
28 in-situ and passively sampling analytes from the gaseous phase. The sorption element was a
29 cylindrical container made of stainless steel net, packed with a carbon sorbent bed –
30 graphitised charcoal *Carbograph 4 35/50* mesh. The recommended working/exposure time of
31 the constructed passive device was determined by carrying out model tests in the laboratory.
32 In addition, a preliminary study was conducted to determine the rate of the emission flux of
33 selected monoterpenes released from the surface of wood-based indoor materials (laminated
34 chipboard) used in residential areas.
35
36
37
38
39
40
41
42

43 25 **Key words:** passive flux sampler, emission flux, monoterpenes, indoor materials, in-situ
44 measurements
45
46
47 27

48 28 **1. Introduction**
49
50 29

51 30 The scientific literature published over the last decade shows the emission of various types of
52 organic compounds (from VOCs and SVOCs group) from furnishings and finishing materials,
53 and this is one of the main indoor processes having a significant impact on the quality of
54 indoor air. VOC and SVOC chemical compounds present in the indoor air may negatively
55 affect the health and comfort of the inhabitants [1-3].
56
57
58
59
60

1 An overview of literature data indicates that emission rates and emission profiles are usually
2 analysed with stationary environmental chambers (small or large scale). Despite the fact that
3 their application in analytical laboratories is costly, time consuming and requires
4 transportation of the tested material into the laboratory, their great advantage is the ability to
5 obtain accurate and reliable analytical information about the type and amount of organic
6 compounds released from the tested material used indoors [4].

7 However, for the individual users of residential areas the transportation of furnishings to the
8 analytical laboratory is an inconvenient and unacceptable option. Moreover, in most cases, the
9 emission measurements requires partial destruction or damage of furnishing material (through
10 grinding or trimming it) to fit it to the size of the stationary environmental chamber. The
11 optimal solution may be a portable device working in a nondestructive pattern for the studied
12 object and that can be used in-situ in the residential area. An additional advantage of such
13 devices would be the ability to work in a passive mode of sampling. Moreover, from an
14 economic point of view, it is important that this type of device does not require an external
15 power supply and is easy to use and cheap in operation. For this reason, an increasing number
16 of analytical laboratories design and construct different types of small-sized portable emission
17 chambers combining the classic features of the miniature emission chamber with the
18 advantages of passive sampling of analytes from the gaseous phase [5]. This type of passive
19 analytical device allows the obtaining of preliminary analytical information on the total
20 amount of organic compounds (TVOC) emitted from the surface of a studied material, or rate
21 of the emission flux of individual compounds, or groups of organic compounds such as
22 formaldehyde [6, 7], toluene, ethylbenzene, xylenes, styrene [8], acetaldehyde, acetone, n-
23 hexanal [9], organophosphate flame retardants [10], phthalate esters [11] and monoterpenes
24 [12]. The general operating principle of small-sized portable emission chambers passively
25 sampling analytes from the gaseous phase is based on the Fick's first law of diffusion. The
26 emission flux largely depends on the organic compound diffusion coefficient in the gaseous
27 phase, the surface area of the studied material, defined by the device's inner dimensions and
28 the passive flux sampler exposure time [4, 5, 13]. To estimate the emission flux of organic
29 compounds from solid indoor materials the transport of organic compounds is mostly
30 described based on the macroscopic scale of the emission process. In a such case the indoor
31 material is considered as a solid body, in which the organic compounds are present in gaseous
32 phase, in pores of the of indoor material [12]. Solid-phase diffusion (organic compounds in
33 adsorbed phase) is negligible compared with the much faster diffusion through pore.

1 The paper presents the characteristic of the new type of analytical equipment for in-situ
2 emission measurements - the home-made in-situ passive flux sampler (PFS) representing a
3 small-scale emission chambers passively sampling analytes from the gaseous phase. The
4 recommended working/exposure time of the constructed analytical device was determined by
5 conducting preliminary model studies in the laboratory for α -Pinene. The selection of α -
6 pinene as a representative monoterpene was based on two factors: the common presence of
7 the α -pinene in wood-based furnishing materials and its relatively low reactivity with other
8 compounds present indoors as compared to the reactivity of other monoterpenes, e.g. D-
9 Limonene [12, 14].

10 Thus, the obtained information allowed for on-site (indoor) preliminary studies to obtain
11 analytical data about the rate of the emission flux of selected monoterpenes (α -pinene, 3-
12 carene and D-limonene) released from the surface of wood-based indoor materials (chipboard
13 laminated on both sides) used in residential areas.

14 15 **2. Materials and Methods**

16 17 **2.1 Design of in-situ passive flux sampler**

18
19 The in-situ PFS used in screening studies to estimate the emissions flux of monoterpenes was
20 constructed from the following components:

- 21 • a cover and a body made of polished stainless steel with a total weight of approx.
22 0.40 kg,
- 23 • seal made of polytetrafluoroethylene (PTFE), providing passive flux sampler
24 tightness during the measurement and to preventing damage to the sorption tube,
- 25 • a commercially available sorption tube (length - approx. 60 mm, external diameter -
26 4.8 mm) made of stainless steel net filled with graphitised charcoal (*Carbograph 4* -
27 mass 300 ± 10 mg of 35/50 mesh).

28
29 The tube filled with a sorbent bed was installed inside the home-made in-situ PFS at a height
30 of approx. 2.5 cm from the surface of studied material. The use of the home-made in-situ PFS
31 in the screening study allows the estimation of the rate of the emissions flux of selected
32 monoterpenes from a flat surface of 25.5 cm^2 . The height of the described device is 3.5 cm.
33 Figure 1 shows a general view of the designed and constructed home-made in-situ PFS.

2.2 Determination of working parameters of home-made in-situ passive flux sampler

The recommended exposure time of the home-made in-situ PFS was defined based on the relationship between the mass of the analyte trapped on a sorbent bed *Carbograph 4*, and the exposure/working time of the passive device. The recommended working time means the time needed to collect (by home made in-situ PFS) the whole amount of the analyte being in the gaseous phase inside the device.

1 µl of freshly prepared standard solution containing 858 ng of α-pinene was applied onto the watch glass placed on a Ø100 mm glass Petri dish. Then, the home-made in-situ PFS with a clean sorption tube was placed on a Petri dish. This experiment was repeated several times for different time periods. Each time the amount of α-pinene retained on a sorbent bed was determined chromatographically using the TD-GC-MS system.

Model studies allowing to determine the home-made in-situ PFS exposure/working time were carried out in two periods of time: in short intervals, i.e. 15, 30, 45, 60 minutes, and long intervals, i.e.: 120, 180, 240, 300 minutes. Three independent replicates of the experiment were performed for each time interval. Each time before carrying out the experiment, the surface of Petri dishes and the inside of home-made in-situ PFS was washed with distilled water and detergent to eliminate the wall memory effect. Additionally the sorption tube with *Carbograph 4* was conditioned for 30 minutes at 300°C

2.3 Thermal desorption and gas chromatography/mass spectrometry analysis parameters

The tubes filled with *Carbograph 4* were subsequently thermally desorbed (*Unity Markes International, Limited Version 2.00, South Queensferry, UK*) for 20 min at 280°C with a high purity helium flow rate of 55 ml/min. As a cold trap (cooled down to 0°C) the quartz tube filled with *Tenax TA* and *Carbotrap* dedicated for collecting VOCs samples in the C_{4/5} to C_{30/32} range was applied. Thermal desorption of the cold trap was performed at 280°C for 5 min with a high purity helium flow rate of 1.0 ml/min. The injection split ratio was set to 1:7. The transfer line temperature was kept at 150°C. Desorbed analytes were analyzed using a gas chromatograph (*Agilent Technologies 6890*) coupled with a mass spectrometer (*5873 Network Mass Selective Detector, working mode - SCAN*). The following gas chromatography

1 parameters were selected: capillary column (*J&W Scientific, Folsom, CA, USA*) DB-5ms (30
2 m × 0.25 mm × 0.25 μm), carrier gas – high purity helium with a flow rate of 1.0 ml/min, GC
3 oven programme: 50°C for 5 min, raised at a rate of 7°C/min up to 150°C, increased at the
4 rate 15°C/min up to 250°C, and finally maintained for 10 min at 250°C. The mass
5 spectrometer ionization source temperature was set to 250°C and the quadrupole temperature
6 was 150°C.

8 **2.4 Chemical standards and calibration of TD-GC-MS system**

10 External standard calibration was carried out with the following monoterpene compounds:
11 (-)-alpha-Pinene (≥99%, δ= 0.858 g/ml), (+)-3-Carene (≥99%, δ = 0.864 g/ml), (S)-(-)-
12 Limonene (≥99%, δ= 0.844 g/ml), Sigma Aldrich, Germany. The liquid solutions were
13 prepared in 10 ml volume glass flasks in two concentration ranges: from 10 to 300 μg/ml and
14 from 300 to 3000 μg/ml. Standard liquid solutions were freshly prepared before the
15 calibration process. Two five-point calibration curves for selected compounds were created:
16 for concentration range from 10 to 300 μg/ml (the coefficient of determination was from
17 0.969 to 0.992); for concentration range from 300 to 3000 μg/ml (the coefficient of
18 determination was from 0.998 to 0.999). Further information on the calibration procedure can
19 be found elsewhere [12, 15]. The limit of detection (LOD) values were from 1.6 ng for α-
20 Pinene to 1.9 ng for D-Limonene.

22 **2.5 Preliminary field studies**

24 The surfaces of two chipboards (with dimensions 45 cm x 30 cm) laminated on both sides
25 were the subject of screening research (field studies) to estimate the rate of the emission flux
26 of selected monoterpenes using the home-made in-situ passive flux sampler. Two chipboards
27 were wood-based construction materials installed in a residential area, and differed from each
28 other significantly in terms of the usage time in the indoor environment. A chipboard covered
29 with a layer of dark laminate was used in the residential area for 30 months from the date of
30 purchase and installation to the beginning of the research. In contrast, a chipboard covered
31 with a light laminate layer was used for one week from the date of purchase and installation.
32 During the preliminary field studies two ready to use **home-made in-situ PFS** were placed on
33 the surface of each wood-based indoor material. After a pre-defined time of exposure, tubes

1 filled with a sorbent bed were removed from PFS and transferred to a glass container, sealed
2 by a polyethylene plug. Subsequently, containers with the sorbent bed were transported to the
3 laboratory and subjected to liberation, separation, identification and final determination of
4 analytes by the TD-GC-MS system.

5 During the studies (14 days), temperature ($22 \pm 2^\circ\text{C}$) and relative humidity (RH. $50 \pm 5\%$)
6 readings were performed. The total number of analyte samples collected during the
7 measurement (14 days) was 28 for each studied laminated chipboard.

9 3. Results and discussion

11 3.1 Determination of working time and the sampling rate characteristics of the home- 12 made in-situ passive flux sampler

14 The model study was carried out in the laboratory to determine the best (recommended)
15 working time and the usefulness of the home-made in-situ PFS for the on-site screening study.

16 The aim of this study was to determine the performance characteristics of the constructed
17 passive device. Figure 2A shows the results of a laboratory test model, illustrating the
18 relationship between the mass of α -pinene retained on the sorbent bed (*Carbograph 4*) of the
19 home-made in-situ PFS and the exposure time of the constructed passive device.

20 As shown in Figure 2A, during 120 minutes of passive device exposure time, approx. 95% of
21 α -pinene applied onto a watch glass and released into the gaseous phase was adsorbed on the
22 sorbent bed. Moreover, data presented in Figure 2A indicate that from 0 to 60 minutes after
23 the installation of the sampler on a glass surface (Petri dish, $\varnothing = 100$ mm) the amount of α -
24 pinene retained on a sorbent bed is proportional (dependent) to the exposure time of the
25 home-made in-situ PFS. In contrast, after 120 minutes, the organic compound contained in the
26 gaseous phase defined by the dimensions of the passive device (internal volume of approx. 64
27 cm^3) was completely adsorbed and retained on the sorbent bed.

28 The recommended working time of the constructed device, which is 120 minutes, was
29 determined by taking into account parameters such as: the very small internal volume of the
30 home-made in-situ PFS, and the specific characteristics of the sorption bed (adsorbing surface
31 ca. $200 \text{ m}^2 \cdot \text{g}^{-1}$) applied inside the passive device. In the case of the home-made in-situ PFS,
32 the sampling rate of the analyte from the gaseous phase depends on two factors: the diffusion
33 coefficient of the analyte in the gaseous phase, and the analyte concentration gradient inside

1 the PFS: concentration gradient between the surface of the studied material and the surface of
2 the sorption element (driving force for the analyte sampling from the gaseous phase).

3
4 To illustrate the changes in the uptake rate of the analyte (α -pinene) from the gaseous phase
5 inside the home-made in-situ PFS, the relationship between the exposure time of the home-
6 made passive device and the sampling rate of the analyte from the gaseous phase was
7 determined (Figure 2B). The analyte sampling rate was calculated based on the following
8 equation (1):

$$9 \quad S_{gp} = \frac{m}{x_{gp} \cdot t} \quad (1)$$

10 where:

11 S_{gp} – the uptake rate of the analyte from the gaseous phase [ml/min]

12 m – the amount of the analyte retained on a sorbent bed [ng]

13 t – exposure time [min]

14 x_{gp} – the amount of analyte in the gaseous phase inside the home-made in-situ PFS [ng/cm³]

15
16 Increase in the exposure time of the home-made in-situ PFS was associated with a decrease in
17 the uptake rate of the analyte from the gaseous phase (from 2.3 ml/min for 15 minutes
18 exposure to 0.50 ml/min for 120 minutes exposure) (Figure 2B).

19 20 **3.2 The emission flux of selected monoterpenes determined using the home-made in-situ** 21 **passive flux sampler**

22
23 The emission flux of organic compounds depends mainly on the age of the indoor material
24 and its indoor usage time. The newer the indoor material (delivered to the enclosed facility
25 directly or after a short storage at the manufacturer), the higher the amount of organic
26 compounds released into the indoor air. With storage/usage time of indoor furnishings, the
27 amount of organic compounds released from the indoor material to indoor environment
28 decreased [12].

29 The results of the preliminary screening studies aimed to estimate the emissions flux of
30 selected monoterpenes (α -pinene, 3-carene and D-limonene) from two types of chipboards
31 laminated on both sides, which differ of indoor storage/usage time are presented in Table 1.
32 The initial screening study was performed using the home-made in-situ PFS. The emission

1 flux of monoterpenes released from the surface of a wood-based indoor materials was
2 determined using the following equation (2)[12, 13]:

$$3 \quad E_f = \frac{m}{A \cdot t} \quad (2)$$

11 where:

12 E_f – the emissions flux of the determined monoterpenes [$\mu\text{g}/\text{m}^2 \cdot \text{h}^{-1}$],

13 m – the mass of the monoterpene retained on a sorbent bed [μg]

14 t – the working time of the *in-situ* PFS [h]

15 A – the area defined by the dimensions of passive device [m^2].

16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

4 The average value of emission flux of α -pinene and 3-carene, released from the tested chipboard surface covered with black laminate, determined for the entire measuring period (two weeks) was $11.0 \pm 5.4 \mu\text{g}/\text{m}^2 \cdot \text{h}^{-1}$ and $6.1 \pm 3.2 \mu\text{g}/\text{m}^2 \cdot \text{h}^{-1}$ respectively.

5 In the initial screening tests, where the object of the study was chipboard covered with dark laminate, the D-limonene emission flux could be determined in only 4 cases. In other measurement series, the amount of D-limonene retained on the sorbent bed was below the limit of detection (<LOD). Therefore, the average value was not determined for the entire measurement period (14 days). Estimation of the mean value for the entire 14-day measurement period, based on 4 results did not provide reliable analytical information reflecting the actual D-limonene emission flux. Measurements summarized in Table 1 indicate that despite the passage of considerable time (30 months) chipboard covered with black laminate, can still be a source of monoterpenes emissions, especially α -pinene and 3-carene, to indoor air.

6 In the case of chipboard covered with a light laminate and stored/used indoors for a very short time (1 week), the average value of emission flux of α -pinene, 3-carene and D-limonene estimated for the entire measurement period (2 weeks) was $35 \pm 16 \mu\text{g}/\text{m}^2 \cdot \text{h}^{-1}$, $4.3 \pm 3.6 \mu\text{g}/\text{m}^2 \cdot \text{h}^{-1}$ and $2.2 \pm 1.8 \mu\text{g}/\text{m}^2 \cdot \text{h}^{-1}$, respectively. Data summarized in Table 1 also show that the tested chipboard surface covered with a light laminate stored/used indoors for 1 week can be a relevant source of monoterpenes emission, especially α -pinene.

25 4. Conclusions

1 The obtained results of preliminary screening studies carried out using the **home-made in-situ**
2 **PFS** indicate that this device could be used in the future for screening studies to determine the
3 rate of emission flux of selected volatile organic compounds, e.g. BTEX (benzene, toluene,
4 ethylbenzene, xylenes) and aliphatic hydrocarbons (from n-heptane to n-undecane), released
5 from the surface of various indoor materials.

6 Preliminary screening studies carried out on laminated chipboards stored/used for different
7 periods of time demonstrated the suitability of the **in-situ PFS** in measuring the emission flux
8 of monoterpenes released from the tested wood-based materials. Preliminary findings from
9 screening suggest that laminated wood-based materials used indoors for shorter periods of
10 time release greater amounts of monoterpenes than materials used indoors for long periods of
11 time.

12

13 **5. Acknowledgements**

14 The study was funded by the Ministry of Science and Higher Education under the “Iuventus
15 Plus” programme for the years 2015-2017; project no. IP2014 028373

16

17 **6. Conflict of Interest**

18 The authors declare that they have no conflict of interest

19

20 **7. References**

- 21 [1] Kim JA, Kim S, Kim HJ, Kim YS (2011) Evaluation of formaldehyde and VOCs emission
22 factors from paints in a small chamber: The effects of preconditioning time and coating
23 weight. *J Hazard Mater* 187:52-57
- 24 [2] Han KH, Zhang JS, Wargocki P, Knudsen HN, Varshney PK, Guo B (2012) Model-based
25 approach to account for the variation of primary VOC emissions over time in the
26 identification of indoor VOC sources. *Build Environ* 57:403-416
- 27 [3] Gallego E, Roca JF, Perales JF, Guardino X (2013) Experimental evaluation of VOC
28 removal efficiency of a coconut shell activated carbon filter for indoor air quality
29 enhancement. *Build Environ* 67:14-25
- 30 [4] Marć M, Zabiegała B, Namieśnik J (2012) Testing and sampling devices for monitoring
31 volatile and semi-volatile organic compounds in indoor air. *Trends Anal Chem* 32:76-86
- 32 [5] Marć M, Zabiegała B, Namieśnik J (2013) Miniaturized Passive Emission Chambers for
33 In Situ Measurement of Emissions of Volatile Organic Compounds. *Crit Rev Anal Chem*
34 43:55-61

- 1 [6] Plaisance H, Blondel A, Desauziers V, Mocho P (2014) Characteristics of formaldehyde
2 emissions from indoor materials assessed by a method using passive flux sampler
3 measurements. *Build Environ* 73:249-255
- 4 [7] Shinohara N, Fujii M, Yamasaki A, Yanagisawa Y (2007) Passive flux sampler for
5 measurement of formaldehyde emission rates. *Atmos Environ* 41:4018-4028
- 6 [8] Kang DH, Choi DH, Lee SM, Yeo MS, Kim KW (2010) Effect of bake-out on reducing
7 VOC emissions and concentrations in a residential housing unit with a radiant floor heating
8 system. *Build Environ* 45:1816-1825
- 9 [9] Yamashita S, Kume K, Horiike T, Honma N, Fusaya M, Ohura T, Amagai T (2010) A
10 simple method for screening emission sources of carbonyl compounds in indoor air. *J Hazard*
11 *Mater* 178:370-376
- 12 [10] Ni Y, Kumagai K, Yanagisawa Y (2007) Measuring emissions of organophosphate flame
13 retardants using a passive flux sampler. *Atmos Environ* 41:3235-3240
- 14 [11] Fujii M, Shinohara N, Limb A, Otake T, Kumagai K, Yanagisawa Y (2003) A study on
15 emission of phthalate esters from plastic materials using a passive flux sampler. *Atmos*
16 *Environ* 37:5495-5504
- 17 [12] Marć M, Namieśnik J, Zabiegała B (2014) Small-scale passive emission chamber for
18 screening studies on monoterpene emission flux from the surface of wood-based indoor
19 elements. *Sci Total Environ* 481:35-46
- 20 [13] Sekine Y, Toyooka S, Watts SF (2007) Determination of acetaldehyde and acetone
21 emanating from human skin using a passive flux sampler—HPLC system. *J Chromatogr B*
22 859:201-207
- 23 [14] Uhde E, Salthammer T (2007) Impact of reaction products from building materials and
24 furnishings on indoor air quality—a review of recent advances in indoor chemistry. *Atmos*
25 *Environ* 41:3111-3128
- 26 [15] Marć M, Zabiegała B, Namieśnik J (2014) Application of passive sampling technique in
27 monitoring research on quality of atmospheric air in the area of Tczew, Poland. *Inter J*
28 *Environ Anal Chem* 94:151-167

30 **Figure captions**

31 **Figure 1.** The overall view of the designed and constructed **home made in-situ PFS**.

32 **Figure 2.** A graphical representation of the relationship between: A - the amount of α -pinene
33 retained on a sorbent bed of the sorption tube, and the working/exposure time of the **home-**

- 1 made in-situ PFS; B - the uptake rate of α -pinene from the gaseous phase and the
- 2 working/exposure time of the home-made in-situ PFS.

For Peer Review

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

Table 1. The results of the determination of the emission flux of selected monoterpenes using home-made in-situ PFS.

The emissions flux of selected monoterpenes released from the surface of the wood-based element covered on both sides by dark laminate, used in residential area for 30 months						
Parameter	first week of research (n = 14)			second week of research (n = 14)		
	α -Pinene	3-Carene	D-Limonene	α -Pinene	3-Carene	D-Limonene
Average [$\mu\text{g}/\text{m}^2\cdot\text{h}^{-1}$]	17.8	7.0	<LOD	8.3	4.9	<LOD
Standard deviation [$\mu\text{g}/\text{m}^2\cdot\text{h}^{-1}$]	1.9	1.3	<LOD	6.5	3.6	<LOD
Median [$\mu\text{g}/\text{m}^2\cdot\text{h}^{-1}$]	14.7	6.8	<LOD	7.2	4.3	<LOD
Min [$\mu\text{g}/\text{m}^2\cdot\text{h}^{-1}$]	10.1	5.5	<LOD	0.41	1.4	<LOD
Max [$\mu\text{g}/\text{m}^2\cdot\text{h}^{-1}$]	15.3	9.2	<LOD	16.5	9.4	<LOD
The emissions flux of selected monoterpenes released from the surface of the wood-based element covered on both sides by a light laminate, used in residential area for one week.						
Parameter	first week of research (n = 14)			second week of research (n = 14)		
	α -Pinene	3-Carene	D-Limonene	α -Pinene	3-Carene	D-Limonene
Average [$\mu\text{g}/\text{m}^2\cdot\text{h}^{-1}$]	42	5.1	2.4	29	3.4	2.1
Standard deviation [$\mu\text{g}/\text{m}^2\cdot\text{h}^{-1}$]	19	4.4	1.6	11	2.8	2.0
Median [$\mu\text{g}/\text{m}^2\cdot\text{h}^{-1}$]	31	2.7	2.5	28	2.7	1.0
Min [$\mu\text{g}/\text{m}^2\cdot\text{h}^{-1}$]	23	0.72	0.68	17	0.62	0.49
Max [$\mu\text{g}/\text{m}^2\cdot\text{h}^{-1}$]	67	10.9	3.7	47	7.5	5.8

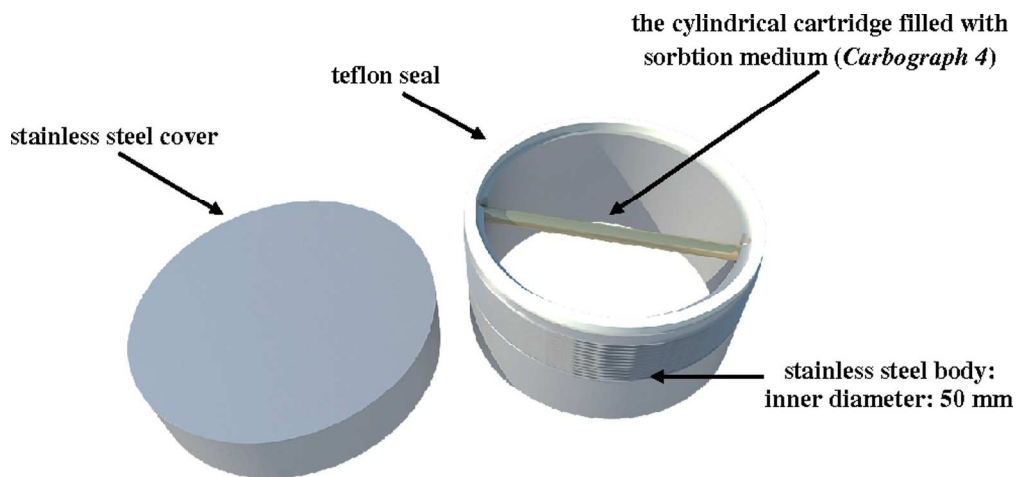


Figure 1. The overall view of the designed and constructed home made in-situ PFS 169x78mm (300 x 300 DPI)

Peer Review

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

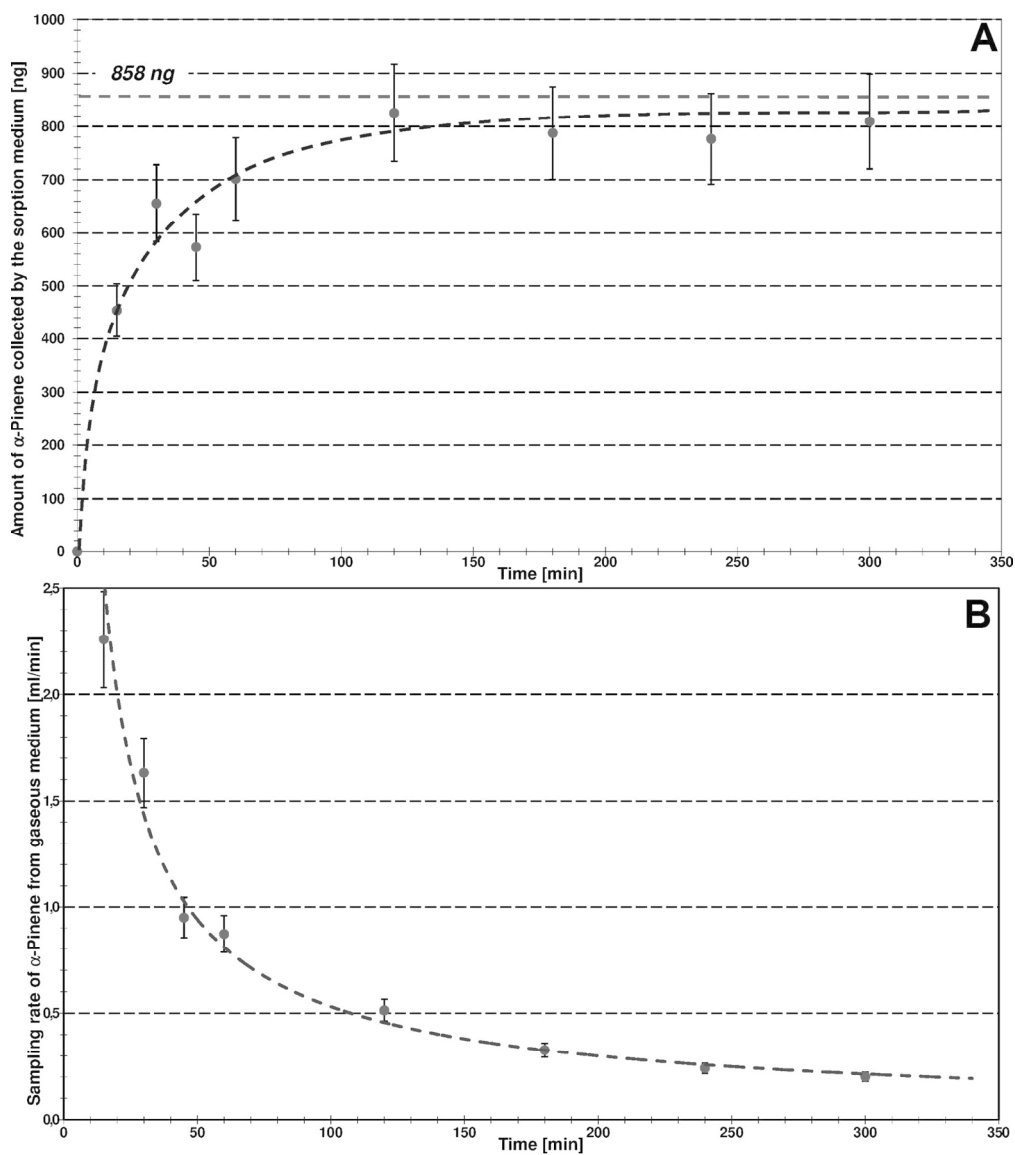


Figure 2. A graphical representation of the relationship between: A - the amount of α -pinene retained on a sorbent bed of the sorption tube, and the working/exposure time of the home-made in-situ PFS; B - the uptake rate of α -pinene from the gaseous phase and the working/exposure time of the home-made in-situ PFS.

191x218mm (300 x 300 DPI)