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7 **Environmental characteristics of a tundra river system in Svalbard. Part 1: bacterial**
8 **abundance, community structure and nutrient levels**

9 Klaudia Kosek^a, Aneta Luczkiewicz^b, Krystyna Koziół^{c,d}, Katarzyna Jankowska^b, Marek
10 Ruman^e, Żaneta Polkowska^{a*}

11 ^aDepartment of Analytical Chemistry, Faculty of Chemistry, Gdansk University of Technology, 11/12
12 Narutowicza St., Gdansk 80-233, Poland

13 ^bDepartment of Water and Waste-Water Technology, Faculty of Civil and Environmental Engineering, Gdansk
14 University of Technology, 11/12 Narutowicza St., Gdansk 80-233, Poland

15 ^cInstitute of Geography, Faculty of Geography and Biology, Pedagogical University in Cracow, Podchorążych 2
16 St., Cracow 30-084, Poland

17 ^dInstitute of Geophysics, Polish Academy of Sciences; 64 Księcia Janusza St., Warsaw 01-452, Poland

18 ^eFaculty of Earth Sciences, University of Silesia, 60 Będzińska St., Sosnowiec 41-200, Poland

19 *corresponding author: zanpolko@pg.edu.pl (Ż. Polkowska)

20

21 **Abstract:** The Arctic hosts a set of unique ecosystems, characterised by extreme
22 environmental conditions and undergoing a rapid change resulting from the average
23 temperature rising. We present a study on an aquatic ecosystem of the Revelva catchment
24 (Spitsbergen), based on samples collected from the lake, river and their tributaries, in the
25 summer of 2016. The landscape variety of the study site and the seasonal change in the
26 hydrological regime modify the availability of nutrients. In general, the upper part of the
27 catchment consists of the mountain rocky slopes which are especially abundant in iron
28 minerals, sulphides and phosphorus minerals. The lower part of the catchment is covered by
29 plants - lichens, saxifrages and bryophytes, which are a different source of nutrients. In the
30 analysed water samples, the maximum concentrations of nutrients such as iron, boron and

31 phosphorus were $0.28 \mu\text{g L}^{-1}$, $4.52 \mu\text{g L}^{-1}$ and $1.91 \mu\text{g L}^{-1}$, respectively, in June, while in
32 September, Fe and B reached the concentrations of $1.32 \mu\text{g L}^{-1}$ and $2.71 \mu\text{g L}^{-1}$, respectively.
33 The concentration of P in September was below the detection limit of $1.00 \mu\text{g L}^{-1}$, which may
34 be explained by the necessity of bacteria to consume it immediately on current needs. We
35 noted also an increase in TOC concentration between the June and September samples, which
36 could originate both from the biomass accumulation in the catchment and the permafrost
37 melting contributing to the hydrological regime of the river. The bacterial community
38 developed in this environment consisted mainly of *Proteobacteria*, *Actinobacteria*,
39 *Bacteroidetes* and *Firmicutes* phylum, while the presence of *Acidobacteria* was less
40 pronounced than in other tundra-related environments. The described catchment shows that
41 despite the relatively small amount of bioavailable nutrients, the Revelva system is biodiverse
42 and one of the most significant biogeochemical changes occurs there in response to seasonally
43 switching water sources.

44 **Keywords:** Arctic, Spitsbergen, Freshwater bacterial community, Bacterial diversity,
45 Nutrients

46 **1. Introduction**

47 The unique Arctic ecosystems, adapted to the extreme environmental conditions of this area,
48 are under pressure due to environmental changes following more than twice as intensive
49 warming of this area as the global average temperature rise (ACIA 2005; AMAP 2017).
50 Rising temperatures affect water supply from shrinking glaciers (Gardner et al. 2013) and
51 permafrost thaw (Frey and McClelland 2009), and they decrease the extent and duration of the
52 snow cover (AMAP 2017), effectively modifying the hydrological regime of the Arctic rivers.
53 The associated landscape changes encompass the exposure of formerly glaciated land and
54 significant shifts in vegetation (Elmendorf et al. 2012; Bjorkman et al. 2018). Another

55 demonstration of such change may be an increased frequency of extreme precipitation events
56 (Łupikasza 2007). The Arctic rivers, draining the catchments changing in these various ways,
57 are predicted to experience a biogeochemical shift towards a groundwater-dominated system,
58 as opposed to one dominated by surface water supply (Frey and McClelland 2009). Since a
59 similar change in supply proportions occurs in the tundra rivers across the summer season
60 (Pulina et al. 1984), we sampled the June and September waters from a Svalbard lake-river
61 system as two biogeochemical composition endmembers, hypothesising their differences will
62 reflect a likely direction in future Arctic river biogeochemistry.

63 A river and its headwaters capture various biogeochemical elements originating from the
64 landscapes it encompasses. The bacterial composition in the Arctic rivers and lakes is linked
65 to the transport of microorganisms from other habitats containing developed microbiota. The
66 changing sources of water supply affect also the pathways of transporting chemical
67 compounds and bacteria into the catchment, which can originate from the airborne pool,
68 glaciers, sea aerosols and permafrost thaw (Houghton et al. 2001; Pomeroy and Wiebe 2001;
69 Hodson et al. 2005, Adams et al. 2010, Kühnel et al. 2013, Górnjak et al. 2016). Although the
70 riverine nutrient concentrations and fluxes in the Arctic in inorganic form are relatively low,
71 such catchments usually discharge high amounts of organic matter (Dittmar and Kattner
72 2003) and the microbial communities harboured by these watercourses may be very diverse
73 (Crump et al. 2012). Carbon, nitrogen and phosphorus can all be limiting nutrients, as related
74 to individual cell physiology and environmental factors (Fagerbakke et al. 1996; Göransson et
75 al. 2011). Nutrient limitation can influence not only elemental ratios in biomass, but also cell
76 volume and shape (Vrede et al. 2002). Phosphorus has been found a common limiting element
77 in the Arctic lakes and ponds, although its enhancing effect on bacterial abundance and
78 production is usually only seen in sites with an increased temperature. A combined effect of

79 phosphorus and organic carbon or nitrogen in water samples may result in increased
80 productivity signals (Graneli et al. 2004; Mindl et al. 2007; Edwards et al. 2014).

81 The main objective of the conducted research was to observe the biogeochemical diversity of
82 the studied aquatic environments with respect to seasonal change, anticipating similar changes
83 with the future shift towards a groundwater-fed system (Frey and McClelland 2009).
84 Furthermore, we investigate whether the nutrients present in the studied catchment are
85 sufficient and available for the development of the bacteria living in it, by studying the
86 interactions between the nutrients, such as phosphate, nitrate, ammonia, or organic carbon,
87 and the bacterial abundance. A background factor influencing them is the variety of
88 hydrological environments and landscapes in this catchment. This information was compared
89 to the quantitative and qualitative data on the local bacterial community composition, showing
90 its variety and adaptation to the environment.

91 **2. Materials and Methods**

92 2.1. Study area

93 The Revelva catchment (Wedel-Jarlsberg Land, southwestern Spitsbergen) is located in the
94 vicinity of the Polish Polar Station Hornsund (77°0'0"N, 15°33'0") (Figure 1). The main river
95 (Revelva) and the lake (Revvatnet) are fed both directly by atmospheric precipitation, snow-
96 fed streams and a river originating from a glacier (Ariebreen), as well as permafrost thaw,
97 especially once the snow has melted. The permafrost thaw in the area is pronounced, as this
98 region is characterised by the highest active layer depth in the Svalbard archipelago,
99 exceeding 2 m (Dolnicki et al. 2013). Furthermore, in the Hornsund station, long-term
100 monitoring (1990-2009) of the active layer temperature at 1 m depth has shown an increasing
101 trend (Dolnicki et al. 2013), leading to the conclusion that permafrost degradation was
102 advancing in that period.

103 Revelva's estuary drains into the bay of Ariebukta. In the upper part of the catchment, the
104 tributary streams originate from rocky mountain slopes. A series of three lakes occupies the
105 valley bottom and contributes to the hydrological diversity of this site. The catchment is
106 characterised by an asymmetry, with a predominance of left tributaries, of which the largest is
107 the proglacial Ariebekken. The sampling was performed mainly on the left side of the river
108 and lake, to reflect this asymmetry in water input and the influence of the lush tundra
109 vegetation in the valley bottom, visited by birds and reindeer herds – biological vectors of
110 chemical species. A further characterisation of this catchment is provided in former
111 publications of our research group (Kozak et al. 2016; Kosek et al. 2018).

112 2.2. Sampling

113 Water sampling in the Revelva catchment was repeated in June and September 2016 in 14
114 locations representing the distribution of water inflows and chemical species into the Revelva
115 system (Figure 1), which were directly comparable to our former study sites (Kosek et al.
116 2018). The choice of sampling months reflected the timespan of the vegetation season and a
117 change in hydrological regime in the area. In June, snow cover is melting and on the sampling
118 date residual snow patches remained in the valley bottom, hence the Revelva system was still
119 influenced chemically and biologically by snowmelt. In September, there occurs an increase
120 in atmospheric precipitation and permafrost thaw, while vegetation season is at an end.

121 **Figure 1.** Location of the studied area in Svalbard and the sampling points in the Revelva
122 catchment.

123 Freshwater samples were collected manually from the Revvatnet (lake) and the Revelva
124 (river) at a distance of 1.5 m from the shore with no headspace into air-tight, chemically clean
125 1L bottles (daily blank sample confirming the purity of the procedure). Pre-cleaning
126 procedure for the bottles included week-long soaking with Milli-Q deionised water and

127 removing the water from the sampling containers several times. The running water was taken
128 from the main stream at depths 20-50 cm below water level. For microbiological analysis,
129 separate sub-surface 50 mL samples were taken and preserved by injecting formaldehyde
130 solution (2% final concentration), then stored at 4°C. An aliquot was taken from the 1 L
131 chemical sample and stored frozen for nutrient analysis (the remaining volume was stored for
132 polycyclic aromatic hydrocarbons analysis – see Part 2, Kosek et al. accepted). The
133 metagenomics samples of 1.5 L were frozen and maintained under such conditions until
134 analysis.

135 2.3. Chemical Analysis

136 All technical specifications of the analytical equipment and methods, including basic
137 validation parameters of the analytical procedures, are given in Table 1. The basic parameters
138 of electrical conductivity (EC) and pH were measured immediately upon return from the field.
139 The concentrations of the following inorganic ions: Li^+ , Na^+ , K^+ , NH_4^+ , Mg^{2+} , Ca^{2+} , F^- , Cl^- ,
140 Br^- , NO_2^- , NO_3^- , PO_4^{3-} and SO_4^{2-} were determined with the use of ion chromatography
141 technique. Phosphorus concentration was also determined in elemental form, as were the
142 concentrations of iron and boron, all with the use of and ICP-MS (Inductively Coupled Plasma
143 Mass Spectrometer). The element concentration CVs of the obtained triplicate results ranged
144 from 0.5 to 1.5%. Carbon, in all organic forms, was measured as non-purgeable organic carbon with
145 a Total Organic Carbon Analyzer TOC-V_{CSH/CSN}, (Shimadzu, Japan) method of catalytic combustion
146 (oxidation) with the application of the NDIR detector. All blanks were prepared with Milli-Q
147 deionised water.

148 **Table 1.** Validation parameters and technical specifications used in the applied analytical
149 procedures.

150 2.4. Quality Assurance / Quality Control (QA/QC)

151 The analytical procedures used to determine individual components in the studied samples
152 have been validated against certified reference materials (CRMs) concordant with ISO Guide
153 34:2009 and ISO/IEC 17025:2005. The data obtained here were subject to strict QC
154 procedures. Prior to pH measurements, a three-point calibration of the electrode was
155 performed with temperature compensation, using MERCK Millipore Certipur® buffer
156 solutions of pH 4.00, 7.00 and 9.00 (25°C). The analysis of elemental nutrients involved the
157 application of Standard Reference Material (RM) NIST 1643e Trace Elements in Water, and
158 RM Enviro MAT ES-L-2CRM, ES-H-2 CRM SCP SCIENCE. The calibration of the
159 apparatus was based on RMs by Inorganic ventures ANALITYK: CCS-4, CCS-6, CCS-1, IV-
160 ICPMS-71A. Potassium hydrogen phthalate by NacalaiTesque (Japan) was used for the
161 calibration of the TOC Analyser. The sensitivity of the applied methods was tested by
162 injecting standard mixtures of the analytes in the measured concentration range. Linear
163 calibration curves of the peak area against standard concentration showed correlation
164 coefficients (R^2) in the range of 0.898–0.999 for all standards. Each sample was analysed in
165 triplicate. The instrumental background was checked by inserting Milli-Q water blanks once
166 per every six samples.

167 2.5. Bacterial Abundance Analysis

168 For the determination of total bacterial number, average bacterial cell volume and bacterial
169 biomass, the collected water samples have been stained with DAPI (4,6-diamidino-2-phenyl-
170 indol) in a final concentration of $2 \mu\text{g mL}^{-1}$ and filtered through a polycarbonate membrane
171 filter with a pore diameter of $0.2 \mu\text{m}$. The samples prepared for bacteria detection have been
172 analysed using the epifluorescence microscope Nikon Microscope 80i with NIS-Elements BR
173 3.0 and MultiScan automated image analysis system. The analysis was carried out using
174 appropriate excitation filters adapted to the used fluorochrome. The total useful microscope
175 magnification was 1200. The image analysis system consisted of a snap-in to the microscope

176 Epifania Mda monochrome high resolution color digital camera (Nikon DS-5Mc-U2).
177 Structure indicators of bacteriocenosis were estimated based on the results obtained in 20
178 consecutive fields of view.

179 2.6. Bacterial Community Structure Analysis

180 The water samples were filtered through sterile 0.2- μ m membrane filters. The total genomic
181 DNA was extracted using the commercially available Sherlock AX kit (A&A Biotechnology,
182 Poland). The membrane filters were first transferred into microcentrifuge tubes containing 0.5
183 g of 0.5 mm zirconia beads and supplemented with 300 μ l of sterile water, 300 μ l of L 1.4
184 buffer and 20 μ l of proteinase K. Next, the samples were placed in a Beadbeater for 60 s. The
185 isolation protocol was then followed according to the manufacturer's instructions. The DNA
186 concentrations of the samples were determined using a ND-1000 UV-Vis spectrophotometer.
187 The extracted DNA was stored at 4°C. The microbial community in the tested samples was
188 analysed using high-speed multiplexed 16S microbial sequencing on a MiSeq platform
189 (Illumina). The microbial community was analysed using the hypervariable regions V3-V4 of
190 the 16S rRNA, regarded as the most appropriate for the Illumina sequencing (Klindworth et
191 al. 2013). The region was amplified using the following primer set: 341F –
192 CCTACGGGNGGCWGCAG and 785R – GACTACHVGGGTATCTAATCC. PCR was
193 conducted using Q5 Hot Start High Fidelity 2X Master Mix (New England Biolabs, Ipswich,
194 MA, USA). Each library was prepared with a two-step PCR protocol based on Illumina's '16S
195 metagenomic library prep guide'. Paired-end (PE, 2 \times 250 nt) sequencing with a 5% PhiX
196 spike-in was performed with an Illumina MiSeq (MiSeq Reagent kit v2) at Genomed
197 (Warsaw, Poland) following the manufacturer's run protocols (Illumina, Inc., San Diego, CA,
198 USA). The primary automatic analysis and the de-multiplexing of the raw sequences were
199 performed with MiSeq, with the use of MiSeq Reporter (MSR) v. 2.6 (BaseSpace). Next
200 sequences were analysed using the bioinformatics pipeline Qiime (Quantitative Insights Into

201 Microbial Ecology) v. 1.8.0. Raw paired-end reads were subjected to the following process:
202 (1) searching and removing both forward and reverse primer sequences using CutAdapt, with
203 no mismatches allowed in the primer sequences, (2) the removal of the low quality sequences
204 not having an average quality of 20 over a 30 bp sliding window based on the *phred* algorithm
205 and a 97% overlap identity, (3) quality-filtered reads were merged based on the overlap of PE
206 read with the use of *fastq-joint*, (4) the sequence reads were classified into OTUs (Operational
207 Taxonomic Units) on the basis of sequence similarity using the UCLUST algorithm, (5) the
208 chimera sequences were detected and removed using the Chimera Slayer algorithm, (6)
209 clustering of operational taxonomic units (OTUs) was performed at 97% similarity using the
210 *uclust* method, based on GreenGenes v. 13.8 database, (7) additionally, samples were
211 hierarchically clustered using Unweighted-Pair Group Method with Arithmetic mean
212 (UPGMA). It is important to note that the software used here limits the identification of
213 taxonomical level to the lowest unequivocally assigned one, i.e. to family if genus and species
214 cannot be recognised. Based on clusters, the diversity indices were estimated, including the
215 Chao1, Shannon, and Simpson indices.

216 2.7. Principal Component Analysis (PCA)

217 Principal Component Analysis (PCA) is a multivariate statistical analysis that allows
218 revealing internal relations in the data set. PCA finds linear combinations of the original
219 variables, referred to as principal components, which provide better descriptors of the data
220 pattern than the original (chemical or physical) measurements and account for most of the
221 dataset variation. The PCA for this study was performed using R v. 3.4.4, using the *prcomp*
222 function, on a log-transformed dataset, except the pH value which is a logarithm.

223 3. Results

224 3.1. The chemical composition of freshwater samples

225 3.1.1. Electrical conductivity (EC), pH and total organic carbon (TOC) concentration

226 The EC values in the collected samples ranged from $34.8 \mu\text{S cm}^{-1}$ to $102.1 \mu\text{S cm}^{-1}$ in June
227 2016, and from $76.9 \mu\text{S cm}^{-1}$ to $174.5 \mu\text{S cm}^{-1}$ in September 2016 (Figure 2a), while pH
228 ranged from 7.0 to 8.0 in both months (data shown in the Part 2 of this article, Kosek et al.
229 accepted).

230 Figure 2b shows that the concentrations of total organic carbon (TOC) in September 2016
231 were higher than in June at all locations. The maximum value of 2.06 mg L^{-1} was measured in
232 September at the R12 site, in the river estuary, while the lowest concentrations occurred in
233 June in the upper part of the catchment (R2-R4 sites); similar concentration was found also at
234 site R9 in June. The smallest seasonal difference in TOC measurements was found at site
235 R13.

236 **Figure 2.** Concentration levels of electrical conductivity and total organic carbon determined
237 in the collected freshwater samples, compared between the studied periods; a) electrical
238 conductivity (EC), b) total organic carbon (TOC).

239 3.1.2. Inorganic ions

240 Figure 3 shows the percentage of total anion and cation concentrations detected in the
241 collected samples. Chloride and sulfates dominate the anion composition both in June and
242 September 2016. In the collected freshwater samples, Cl^- constituted almost 46% of all
243 detected ions both in June and September 2016. Sodium and calcium were predominant
244 cations in all samples except sites R4, R9 and R14 in June, when magnesium exceeded
245 calcium concentrations. A marked change in the cation composition occurred from June to
246 September, with calcium becoming the most abundant cation in all September samples.

247 Nitrogen occurred in the Revelva catchment in ionic form, especially as NH_4^+ and NO_3^- .
248 Nitrate occurred at concentrations ranging from 1.27 to 3.22 mg L^{-1} , with an increase in
249 September (June median concentration amounted to 1.55 mg L^{-1} , while September median
250 equaled 1.99 mg L^{-1}). Ammonium concentrations spanned 0.03 – 0.43 mg L^{-1} , with somewhat
251 higher concentrations in June (June median = 0.11 mg L^{-1} , September median = 0.08 mg L^{-1}).
252 Of other ionic nutrients, neither nitrite nor phosphate was detected, which implies their
253 concentrations were below 0.06 mg L^{-1} .

254 **Figure 3.** Percentage anion and cation composition of the collected samples.

255 3.1.3. Elemental nutrients

256 Among nutrients, elemental B, P and Fe were analysed here in this speciation form (Table 2).
257 The maxima in both boron and phosphorus concentrations occurred in June, while iron
258 concentrations exhibited a marked increase in September.

259 **Table 2.** Concentrations (\pm standard deviation, SD) of elemental nutrients in the collected
260 freshwater samples.

261 3.2. Microbial Community

262 In the collected freshwater samples, the highest bacterial biomass (BB) was detected in
263 September 2016, in the sampling point R5, at 8.47 $\mu\text{g C L}^{-1}$. BB was strongly linked to total
264 bacterial number (TBN) which at this point was also the biggest ($42.1 \cdot 10^4 \text{ cell mL}^{-1}$). Figure
265 4 presents the TBN, BB and average cell volume (ACV) detected in freshwater samples in
266 both months. Notably, the ACV increased in areas where the number of bacteria was lower in
267 both June and September.

268 **Figure 4.** Comparison of bacterial abundance (total number), average bacterial cell volume
269 and bacterial biomass in the Revelva catchment in June and September 2016.



270 3.3. Bacterial Taxonomy

271 In the three tested points (R4, R8 and R14) at two occasions (in June and September), 840348
272 sequences (reads) were detected. Among them, 805 000 were linked to the bacterial and 708
273 to the archaeal domain, while 34640 were not identified (not found in the conventional
274 databases). Samples R4 and R14 showed the highest bacterial diversity, while samples R8
275 were the least diverse (Table 3), regardless of the period of sampling.

276 **Table 3.** Number of reads and OTUs as well as species richness estimate (Chao1) and
277 diversity indices (Shannon and Simpson) for the sampling points.

278 A cluster analysis of sequence data for the examined samples and bacterial community
279 structures with relative abundances at the phylum level (based on the number of Illumina
280 MiSeq-based method) are given in Figure 5 and Figure 6. The predominant bacterial phyla
281 found in the studied catchment were: *Proteobacteria*, *Actinobacteria*, *Bacteroidetes*,
282 *Firmicutes* and *Planctomycetes*; *Verrucomicrobia*, *Tenericutes*, *Cyanobacteria* and
283 *Acidobacteria* were also identified.

284 **Figure 5.** Cluster analysis of bacterial community structures.

285 **Figure 6.** Bacterial community structures and relative abundances based on the number of
286 reads (a) and OTUs (b) for major phyla (>1%) identified in tested water samples in June and
287 September 2016.

288 The bacterial taxonomic composition at the family level is presented in Table 4. We found the
289 ammonia oxidizing bacteria (AOB) and archaea (AOA), as well as the nitrite-oxidizing
290 bacteria (NOB) in each sampling point, at a total concentration below 0.5% of the total reads.
291 The identified AOA were mainly represented by *Nitrosopumilus* and *Candidatus*
292 *Nitrososphaera*, while among AOB, *Nitrosococcus* from *Gammaproteobacteria* class, and

293 *Nitrospira*, *Nitrosovibrio* from *Betaproteobacteria* class were detected. The nitrite-
294 oxidizing taxa (*Nitrospira* and *Nitrobacter*) in the tested water samples were at a very low
295 level (<0.08%).

296 **Table 4.** Family level taxonomic composition in the Revelva catchment (among 274 families
297 reported in this study, first 40 are presented).

298 Within the predominant (at the study site) *Proteobacteria* phylum, *Alphaproteobacteria*
299 constituted 8.83% – 18.57% of total reads, and were mainly represented by genus
300 *Thalassospira* (*Rhodospirillaceae* family), which was reported as involved in the phosphorus
301 cycling in nutrient-limited environments (Hütz et al. 2011). In this study, *Thalassospira* was
302 mainly reported in point R8 (R8-J – 5.31% and R8-S – 13.45%) and R14 (R14-J – 4.76% and
303 R14-S – 3.79%), influenced by the tundra soil active-layer controls, while in R4 it reached
304 less than 0.5%.

305 Among the *Betaproteobacteria* class (*Proteobacteria*), the predominant genera were from
306 *Comamonadaceae* family: *Polaromonas* (from 1.8% to 5.1%), *Rhodoferax* (from 1.6% to
307 5.6%), from *Oxalobacteraceae* family: *Polynucleobacter* (from 0.2% to 3.4%) and
308 *Herminiimonas* (from <0.1% to 2.5%). The *Rhodoferax* genus was represented in this study
309 mainly by *R. ferrireducens* sp. nov., a facultatively anaerobic bacterium that oxidizes acetate
310 with the reduction of Fe (III) (Finneran et al. 2003). Another dissimilatory iron reducing
311 bacteria, *Geobacter*, was also found in the studied river-lake system (at abundances up to
312 1.7%). This is a mesophilic bacteria from the *Geobacteraceae* family, class
313 *Deltaproteobacteria*. Sulfate-reducing bacteria were also detected in this study, e.g.
314 *Desulfovibrio* spp. from *Deltaproteobacteria* (up to 0.3%).

315 In *Bacteroidetes* phylum, the *Flavobacteriaceae* family formed from 2.24% to 5.57% of total
316 reads, represented mainly by genus *Flavobacterium* (from 1.55% to 4.54%). *Bacteroidetes*



317 phylum was also represented by *Sphingobacteriaceae* family (from 0.31% to 2.54%), as well
318 as by *Flexibacteraceae* family (from 0.55% to 2.22%). Interestingly, a higher abundance of
319 *Bacteroidetes* phylum was noted in the sampling point R4 (up to 11.8%), when compared
320 with R8 (up to 8.0%) and R14 (up to 7.7%), while in the case of the *Actinobacteria* phylum,
321 an opposite pattern was found (R4 - up to 11.97%; R8 - up to 22.9% and R14 - up to 15.39%).

322 4. Discussion

323 4.1. The chemical composition of freshwater samples

324 4.1.1. Electrical conductivity (EC), pH and total organic carbon (TOC) concentration

325 The noted EC and pH values do not deviate significantly from the former measurements in
326 hydrochemical studies of the Hornsund fjord area (including the Revelva catchment),
327 although the area is characterised by a marked hydrochemical variability. For example, all
328 samples collected in the previous years in the Revelva catchment, as well as those collected
329 this study, were characterised by a near-neutral pH (Ruman et al. 2012; Kozak et al. 2016;
330 Kosek et al. 2018). They also resembled in this respect a nearby lake-stream system in the
331 Brategg Valley (Górniak et al., 2016), however in the Revelva catchment the EC was higher,
332 approximately doubling the values noted by Górniak et al. (2016).

333 The TOC concentrations in this fluvial system showed a spatial pattern of higher values in the
334 lower part of the river system, indicating the likely transport of TOC downstream and its
335 accumulation from the biological production in the lakes and the surrounding tundra. Such
336 spatial distribution was especially visible in the beginning of the season, when the upper parts
337 of the catchment were still partly snow-covered. The maximum TOC value in Revelva
338 approximately doubled the maximum DOC (dissolved organic carbon) value noted in the
339 biggest lake of the Brategg Valley (in the first half of August, Górniak et al., 2016). We noted
340 also a temporal increase in TOC concentration between the June and September samples,

341 which could originate from the biomass accumulation in the catchment, but also from the
342 permafrost melting contributing to the hydrological regime of the river. The point in which
343 such a change was the least notable (R13) was fed by glacial meltwater.

344 4.1.2. Inorganic ions

345 The high concentrations of chloride and sodium in the collected samples testify to the
346 important influence of sea spray on the local precipitation (Kosek et al. 2018), which feeds the
347 surface waters, especially as snow melt in June. Another important contributor to the ion
348 composition of the Revelva system waters is rock weathering, which increases the
349 concentrations of calcium, magnesium and potassium ions. It is the most likely source of the
350 predominant concentration of Ca^{2+} in September, when groundwater related to permafrost
351 thaw feeds the surface waters in a significant proportion (McKenzie and Voss 2013;
352 Szumińska et al. 2018).

353 The content of ammonium and nitrate ions in aquatic environment is an important factor in
354 the development of microorganisms, especially in low-nutrient environment (Rivkina et al.
355 2000), and the interesting fact found for the Revelva catchment was their reverse pattern of
356 seasonal change in concentration. McNamara et al. (2008) reported similar temporal patterns
357 of NH_4^+ and NO_3^- concentrations in the Kuparuk river system in Alaska, connecting them to
358 the origin of ammonium from snowmelt (including leaching the top layer of soil by snowmelt,
359 which in anoxic conditions produces more NH_4^+). Such a mechanism is corroborated by the
360 finding of ammonia oxidising archaea and bacteria in the microbial population of the Revelva
361 system. As McNamara et al. (2008) point out, the following increase in nitrate concentration
362 could be a result of nitrification occurring in the well-mixed stream waters. Also in this study,
363 the microbial communities responsible for nitrogen transformation were detected, which was
364 described in details in points: 3.2 and 4.2.



365 4.1.3. Elemental nutrients

366 The nutrients B, Fe and P in the studied catchment waters may originate from local rock
367 weathering, however B was also found to occur at higher concentrations in precipitation than
368 in surface waters, in the neighbouring Fuglebekken catchment (Kozak et al. 2015).
369 Geologically, the studied part of Spitsbergen is built of Proterozoic crystalline rocks that in
370 the coastal zone of the valley are covered by Quaternary clastic formations. The crystalline
371 bedrock is formed of various kinds of metamorphic rocks, mainly gneisses, mica-schists,
372 quartzites, migmatites, marbles, amphibolites and calcareous-silicate rocks (Marszałek and
373 Wąsik 2013). These rocks are characterised by various degrees of fissuring and they are
374 markedly weathered in the upper parts of the catchment. The river valleys are filled with
375 coarse clastic material, interdigitating with moraine till formations from local glaciers. The
376 coastal zone is covered by coarse gravels and boulders (Marszałek and Wąsik 2013).
377 Throughout the rock formations of the Revelva catchment, ore-bearing mineral veins occur,
378 which are especially abundant in iron minerals: many are ankerite or quartz-ankerite veins,
379 and they contain other iron minerals, such as pyrite, chalcopyrite, pyrrhotite, sometimes also
380 magnetite and haematite (Wojciechowski 1964). Thus the increase in Fe concentrations in
381 September may be caused by the occurrence of groundwater associated with the active layer
382 of permafrost, which gains more importance in the hydrological regime of the Revelva once
383 snow patches disappear in the catchment, and leaches iron from ore-bearing layers. Some of
384 these minerals are sulphides, and these occur on the whole left side of the Revelva, which
385 would contribute to the formation of abundant sulphate in the runoff. In the top part of the
386 catchment (Gangpasset), Smulikowski (1965) mentions also the occurrence of phosphorus
387 minerals (apatite), although this did not raise elemental phosphorus or phosphate
388 concentrations in the studied samples to the detection level. In fact, the only points and
389 sampling occasion when we detected elemental phosphorus was the lowest part of the

390 catchment in June. This could reflect the elevated concentrations of phosphorus-containing
391 particles in snowpack and the correlation of inorganic phosphorus removal with runoff, as
392 was observed in a catchment in Alaska (McNamara et al. 2008). The general pattern matches
393 the low concentration levels of inorganic phosphorus in other Arctic rivers, which tend to
394 carry nitrogen and phosphorus mainly as organic compounds (Dittmar and Kattner 2003).

395 4.2. Microbial Community

396 The parameters such as the TBN, ACV and BB, provide means for a general monitoring of
397 temporal and spatial changes of the bacterial abundance in river-lake systems. In this study,
398 the observed values were lower than in a neighbouring valley. Both the maximum TBN and
399 BB values were slightly less than the minimum values of these parameters noted by Górnjak
400 et al. (2016), factoring in the presence of <4% archaea in their estimations of biomass. As the
401 Revelva system is only glacially fed in a small proportion, while the Bratæg system was a
402 typical proglacial succession sequence, this can reflect the influence of nutrient and cell
403 supply from the glacier, magnified by the increasing temperature downstream, on the Bratæg
404 data (*cf.* Graneli et al. 2004; Mindl et al. 2007). However, the contrast is not very strong. In
405 fact, the values found in Mackenzie river by Vallières et al. (2008), as well as in the Kuparuk
406 river and the Toolik lake by Hobbie et al. (1983) encompassed the range of values found in
407 Revelva, and they were not different from the values characteristic for temperate rivers.

408 What is more interesting, however, in the Revelva catchment, are the spatial patterns observed
409 at a smaller scale. In these, nutrient supply is likely to play a significant role. For example, in
410 September 2016, in sampling points R5, R6 and R10, bacteria were more abundant than in
411 June 2016, which is consistent with the greater availability of NO_3^- and Fe at the time.
412 Furthermore, ACV was higher upstream from TBN maxima, and this disparity between the
413 two indices may be interpreted assuming the organisms to represent various ecological tactics



414 (Golovlev 2001) which depend on the nutrient availability in the catchment. In oligotrophic
415 environments, organisms are less likely to reproduce fast, so the remaining cells may grow in
416 size (Cole et al. 1993; Šimek 1994).

417 4.3. Bacterial Taxonomy

418 The tested points differ significantly in terms of nutrient sources originating from bedrock,
419 local plant tissue, or supplied by animal vectors and water inflow. Points R4 and R8
420 experience low nutrient input from local vegetation, while point R14 is located in a boggy
421 area, rich in cyanobacteria and bryophytes, assisted by lichens and saxifrages in varying
422 proportions (Kumar et al. 2017). Additionally it should be noted that R8 is located at the
423 drainage point of the Revvatnet lake to the river, R14 is located in a small stream on a raised
424 marine terrace, while R4 represents stagnant water of a small lake. These environmental
425 factors corroborated the changes in bacterial biodiversity, which indicated the main river at
426 the lake drainage point to be the least diverse, confirming the generally observed drop in
427 bioersity from headwaters for main watercourses (Crump et al. 2012; Górnjak et al. 2016). It
428 should also be noted that across the summer season the predominant bacterial families have
429 changed in the tested points (Table 4), especially at the lake drainage point, where the
430 abundance of a certain family could change by as much as 8%.

431 The obtained results are in agreement with those previously presented by Ntougias et al.
432 (2016), where members of the *Proteobacteria*, *Actinobacteria*, *Bacteroidetes* and *Firmicutes*
433 were found predominant for the Revelva catchment. Interestingly, members of those phyla
434 were also predominant in the Arctic tundra (Nissinen et al. 2012). Several identified genera
435 are psychrofiles or psychrotolerant (*Rhodoferrax ferrireducens* sp. nov., Finneran et al. 2003),
436 typically found on glacier surfaces (*Polaromonas*, Hell et al. 2013; Gawor et al. 2016) or in
437 Arctic fjord sediments (*Herminiimonas*, Canion et al. 2013). Furthermore, the *Acidobacteria*

438 phylum found here is considered an indicator of the tundra influence (Männistö et al. 2013),
439 although it was detected at a lower relative abundance than is typical for Arctic settings. The
440 *Acidobacteria* phylum was reported as predominant for instance in Canadian, Alaskan, and
441 Siberian Arctic soils (Neufeld and Mohn 2005; Wallenstein et al. 2009; Rawat et al. 2012;
442 Männistö et al. 2013), but not in Kongsfjorden tundra soil, which has pH close to neutral. In
443 Kongsfjorden tundra soil, the dominance of the *Proteobacteria* over the *Acidobacteria* was
444 reported by Tveit et al. (2013). Moreover, water samples taken from the Revelva catchment
445 were slightly alkaline, characterized by pH from 7.1 to 7.9, while the pH growth optima for
446 *Acidobacteria* is in the range from 3 to 6 (Jones et al. 2009). This can explain why, in this
447 study the *Acidobacteria* phylum accounted for a small share of the population from 0.22% to
448 0.6%, while the most abundant were *Proteobacteria* (from 43% up to 53%). In the Arctic
449 rivers and lakes, biochemical carbon cycling may be limited by the availability of N and P. In
450 this study, specialist bacteria were detected, utilising various nitrogen sources. The ammonia
451 oxidizing bacteria (AOB) and archaea (AOA) as well as anammox bacteria, which use
452 ammonia as a substrate for metabolism, occurred in the studied points at concentrations which
453 can substantially influence the experienced nutrient levels. Despite the limited robustness of
454 gene-fragment assignment to a certain species or even genus, it has to be noted that the
455 ammonia-metabolising organisms were very likely represented in the studied catchment. In
456 particular, the abundance of AOB in the tested samples was at a comparable level as obtained
457 in wastewater processes, where besides the higher temperature also ammonia concentrations
458 are several times higher. Despite the nitrite concentration below detection limit in the water
459 samples ($<0.06 \text{ mg L}^{-1}$), the nitrite oxidising bacteria (NOB) were also detected, although at a
460 very low level. Furthermore, *Nitrospira*, besides being an NOB, was reported to convert
461 ammonia directly to nitrate in comammox process (Daims et al., 2015).

462 Another possible metabolic path, where nitrogen serves as both an electron acceptor and an
463 electron donor, is anaerobic ammonium oxidation (anammox); NH_4^+ is oxidized to N_2 gas
464 using NO_2^- (Lotti et al. 2014). In the conducted study, anammox bacteria could be found in
465 the *Brocadiaceae* family. However, the occurrence of anammox bacteria in the studied
466 samples does not confirm their anammox activity, due to the oxygen presence. The detected
467 anammox bacteria can catalyse other oxidation/reduction processes or be transported with
468 runoff from an occasionally deoxygenated area of the boggy biological soil crust.

469 Finally, there could be even nitrogen-fixing bacteria in the studied catchment, as the second-
470 most abundant phylum in terms of read numbers, *Actinobacteria*, includes members linked to
471 the symbiotic nitrogen-fixing associations with plants (Cernava et al. 2015). On the other
472 hand, denitrifying bacteria could be found in *Betaproteobacteria* class (*Proteobacteria*),
473 *Comamonadaceae* family, genus *Herminiimonas*, detected in this study. Such were described
474 before by Canon et al. (2013) in samples from Svalbard fjord sediments. Overall, the
475 bacterial community of this river-lake system has members occupying various niches in the
476 nitrogen cycle.

477 In Arctic aquatic ecosystems, the degradation of permafrost typically increases phosphorus
478 export to surface waters, although it can be consumed immediately on current needs. Thus,
479 the presence of *Rhodospirillaceae* supports the explanation of the provenance of extra
480 phosphorus supplies to maintain the described microbial abundance. Furthermore, the second-
481 most abundant phylum in terms of read numbers, i.e. *Actinobacteria*, includes a large number
482 of taxa exhibiting P solubilization and mineralization ability, which seems to be crucial in
483 Arctic lakes and waters. In this study, mineral forms of phosphorus were mostly below the
484 detection limit of $1.0 \mu\text{g L}^{-1}$, and the only sites with phosphorus detected were located near
485 the river mouth in June. This highlights the low availability of this nutrient once the microbial
486 activity has increased in the summer season.

487 Further families detected in the studied catchment contain species and genera capable of iron
488 and sulphur compound reduction. The *Comamonadaceae* and *Geobacteraceae* family contain
489 iron-reducing genera. The higher concentration of Fe noted in September coincided with the
490 higher abundance of iron reducing bacteria from the *Comamonadaceae* family, but not from
491 the *Geobacteraceae* family. The latter contains species possibly reducing both ferric iron
492 and/or sulphur compounds at a low temperature (Nixon et al. 2017), and other sulfate-
493 reducing bacteria were likely found in the studied catchment, although at low abundances.

494 Further links between the bacterial taxonomy and the utilised sources of organic carbon can
495 be found. The phylum *Planctomycetes*, commonly detected in permafrost-affected soil
496 ecosystems as minor microbial components (Steven et al. 2007; Wagner et al. 2009; Kim et
497 al. 2014; Hultman et al. 2015) and predominant in lichen-covered soil (Ivanova et al. 2016),
498 was represented in abundance at points R8 and R14, which were surrounded by an area
499 covered by mat-forming cyanobacteria, bryophytes, lichens and saxifrages. Together with
500 *Planctomycetaceae* family, bacteria from *Flavobacteriaceae* (*Bacteroidetes* phylum) family
501 are common inhabitants of detrital aggregates, linked to algal bloom and the degradation of
502 algal sulfated polysaccharides (Kolton et al. 2016; Ivanova et al. 2016). The abundance of
503 *Alphaproteobacteria* members in the analysed water samples can also be explained by their
504 participation in degradational and symbiotic relationships with lichens, and also in the
505 nitrogen fixing, since nitrogenases are known to be ubiquitous among endophytes (Grube and
506 Berg 2009). Local wildlife (especially birds, reindeer, and other terrestrial mammals, such as
507 the polar fox and the polar bear) may also act as nutrient vectors (Mindl et al. 2007), in which
508 their gut microbiota play a yet poorly understood role. Few studies have examined the gut
509 microbiome of animals living in the polar environments to date (Glad et al. 2010). In the case
510 of arctic-breeding shorebirds (Grond et al. 2017), gut microbiota were dominated by
511 *Clostridia* and *Gammaproteobacteria*, but the environment of their nesting area was

512 comprised predominantly of *Proteobacteria*, *Firmicutes*, *Actinobacteria*, *Verrucomicrobia*
513 and to a lesser extent of *Bacteroidetes*, the core bacteria noted also in this study. Gut of the
514 wildlife is an important source of organic nitrogen and phosphorus in the oligotrophic
515 conditions of the Arctic. As bioindicators of fecal contamination serve the following bacteria:
516 *Escherichia coli*, *Clostridium perfringens*, members of *Enterococcus* and *Bifidobacterium*
517 genera. In this study, the key fecal indicators were detected in all sampling points:
518 *Enterococcus* spp. by less than 0.01% of total reads, *Escherichia* up to 0.28%, while
519 *Bifidobacterium* up to 1.71% and *Clostridium* up to 2.29%. Among other *Firmicutes*,
520 *Candidatus Phytoplasma*, *Acholeplasma* and *Mycoplasma* were present in each sampling
521 point at up to 0.67%, 1.47% and 1.94%, respectively. Their role and abundance need further
522 study, yet their presence points to the influence of animals supplying nutrients throughout the
523 catchment with faeces, a fact confirmed by the observations of reindeer herds and various bird
524 species in the catchment.

525 4.4. Statistical analysis on the nutrient-dependence of the bacterial abundance in the Revelva 526 catchment

527 An increasing number of studies, e.g. Stibal et al. 2008; Petrone and Richards 2009;
528 Jørgensen et al. 2014; Ntougias et al. 2016, have shown that despite the small amount of
529 bioavailable nutrients, persistent subfreezing temperatures, prolonged darkness during winter,
530 and exposure to sunlight during summer, aquatic bacteria lead a relatively abundant life in the
531 Arctic (Chu et al. 2010). To explore patterns in the correlations between nutrient levels and
532 bacterial abundance, we have conducted principal component analysis (PCA) on a set of
533 chosen variables. In a coordinate system described by the two first principal components,
534 there was a clear division between samples collected in the early and late summer (Figure 7
535 top). Moreover, the clear division between the bacterial communities in these two periods can
536 be read from the performed cluster analysis (Figure 5).

537 This seasonal division was consistent with the higher concentration of TOC in September
538 (likely originating from both the decomposing plant tissue and permafrost thaw), as well as
539 the higher bacterial cell counts (but not higher cell volumes). The organic matter and most
540 ionic concentrations are typically higher in permafrost thaw waters than in melting snowpack,
541 hence the seasonal division in sample chemical composition can be interpreted as a change in
542 hydrological regime over the summer (Pulina et al. 1984). The PCA showed a less distinct
543 division between hydrological environments, however there were variables in each separate
544 season that differentiated lake and flowing water as well (Figure 7 middle and bottom). The
545 PCA demonstrated also that the variability connected to bacterial volume was disconnected
546 from the variability related to the bacterial number, which may represent the application of
547 different ecological tactics in the bacterial community at conditions of nutrient abundance and
548 shortage. In general, the lake samples were more likely to contain bacteria with high cell
549 volume, while the stream and river environments facilitated higher bacterial numbers and
550 most likely also higher biodiversity, as could be observed in the taxonomic characterisation of
551 the selected few samples.

552 **Figure 7.** Principal component analysis results for nutrient concentrations and the bacterial
553 community parameters (top). The two graphs below represent the two studied periods (June in
554 the middle and September at the bottom), which were clearly divided in the analysis of the
555 whole dataset.

556 The seasonal difference in nutrient abundance was clearly depicted by the PCA. Only
557 ammonium and boron showed higher concentrations in June samples than in September. The
558 closely correlated variable groups in the whole dataset were: [Fe]-[TOC], [SO₄²⁻]-[Cl⁻], and
559 [Mg²⁺]-[K⁺]-[NO₃⁻]. However, in June the strongest correlations were found between [Na⁺]-
560 [Cl⁻]-[B] and [Fe]-[SO₄²⁻]-[TOC]. The Na⁺ and Cl⁻ clearly indicate the sea spray source,
561 which is also present in the local precipitation (including snow cover). Their presence in

562 surface drainage may be modified at this time by elution intensity from snowpack. Boron is
563 therefore likely to originate mostly from seawater as well, through elevated concentrations in
564 precipitation (Kozak et al. 2015). Indeed, in several samples the B/Cl⁻ ratio was close to the
565 0.000241 value reported as the mean for North Atlantic and North Pacific water (Lee et al.
566 2010), and the mean ratio has dropped from 0.000210 to 0.000163 (from June to September).
567 Since rock sources normally contain proportionally more B than seawater (Arnórsson and
568 Andrésdóttir 1995), such a drop can be interpreted as an indication of boron being depleted by
569 the local microbial community. The connection between Fe and SO₄²⁻ corresponds well to
570 their common source in pyrite and chalcopyrite decomposition, yet their increased
571 concentrations occur mainly in the downstream part of the catchment ('river' on Figure 7
572 middle), and the simultaneous accumulation of TOC is in agreement with the fact that pyrite
573 decomposition is microbially mediated.

574 In September, the variables with the closest relationship were [Mg²⁺]-[Ca²⁺]-[SO₄²⁻], [B]-
575 [TOC] and [Fe]-[pH]. The first group likely corresponds to rock weathering, and these ions
576 achieved the highest concentrations in the waters supplied from the glacier and near the river
577 mouth. The boron and TOC association, combined with their close correlation to TBN,
578 confirms the likely use of boron in biological processes enabling the bacterial community to
579 grow and release organic substances. Finally, the Fe concentration can be regulated by the pH
580 of the environment, however the direction of the relationship found here is contrary to the one
581 based on solubility of iron (and its speciation forms) only. Potentially, the oxidation-reduction
582 potential of the water and the presence of iron bacteria modify the pattern more significantly
583 here (Hem and Cropper 1962). Such an interpretation is confirmed by the difference between
584 lake and flowing water iron concentrations, with the lowest concentrations found in the
585 Revvatnet (large lake) and some main river samples (below that lake), while the highest
586 values were noted in the headwaters of the upper part of the catchment (data not shown).

587 In conclusion, the catchment chemical state and the abundance of bacteria undergo a notable
588 shift during the summer season, which indicates features of increased groundwater supply, but
589 also an increased microbial activity and resulting nutrient depletion. The parameters with
590 elevated concentrations in September samples are likely candidates for more important
591 biogeochemical factors in the future of the Arctic rivers. However, the bacterial activity may
592 revert some of the typically observed patterns, e.g. by depleting phosphorus and nitrogen in
593 inorganic forms. The complex feedbacks between such processes require further
594 investigations.

595 **5. Final remarks and conclusions**

596 The rapid environmental change in the Arctic is likely to bring complex biogeochemical
597 shifts, some of which can be anticipated by studying changes in a catchment with a seasonally
598 changing water supply. The results obtained here confirm that freshwater environments in the
599 Arctic contain a low amount of bioavailable forms of nutrients (especially phosphorus)
600 needed for bacterial growth, an amount that is altered as the summer season progresses, in
601 connection to switching nutrient sources and microbial activity. This requires applying
602 various ecological tactics to survive (e.g. investing in cell growth or reproduction in different
603 environments / periods). Despite this, a number of bacterial phyla occupy the studied
604 catchment (mainly *Proteobacteria*, *Actinobacteria*, *Bacteroidetes*, *Planctomycetes* and
605 *Firmicutes* by the order of abundance, however by the number of OTUs *Acidobacteria* were
606 at least the third most important phylum in all samples). The determined bacteria were
607 characterised by high biodiversity indices and a development of multiple survival strategies,
608 as well as a variety of metabolic pathways developed to utilise the existing small nutrient
609 concentrations. The community has also shown a remarkable ability to adjust to the existing
610 conditions changing over the summer season, showing a change in taxonomical composition
611 and relative family abundances between the June and September samples (by up to 8%). For

612 future studies, we recommend especially the studies of this complex relationships between the
613 bacterial community and their chemical environment at a higher temporal resolution,
614 combined with the determination of activity of various bacterial groups.

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625 **References**

626 ACIA (2005). Impacts of a warming Arctic: Arctic climate impact assesment.

627 Adams H.E., Crump B.C., Kling G.W. (2010). Temperature controls on aquatic bacterial
628 production and community dynamics in arctic lakes and streams. *Environ. Microbiol.* 12,
629 1319-1333.

630 AMAP (2017). Snow ice and Permafrost in the Arctic (SWIPA) . Oslo, Norway, xiv + 269
631 pp. Arnórsson S., Andrésdóttir A. (1995). Processes controlling the distribution of boron and
632 chlorine in natural waters in Iceland. *Geochimica et Cosmochimica Acta* 59, 4125–4146.

633 Bjorkman A.D. et al. (2018). Plant functional trait change across a warming tundra biome.
634 *Nature* 562, 57–64.

635 Canion A., Prakash O., Green S.J., Jahnke L., Kuypers M.M., Kostka J.E. (2013). Isolation
636 and physiological characterization of psychrophilic denitrifying bacteria from permanently
637 cold Arctic fjord sediments (Svalbard, Norway). *Environ Microbiol.* 15, 1606-18.

638 Cernava T., Aschenbrenner I.A., Grube M., Liebming S., Berg G. (2015). A novel assay for
639 the detection of bioactive volatiles evaluated by screening of lichen-associated bacteria. *Front*
640 *Microbiol.* May 1, 6-398. doi: 10.3389/fmicb.2015.00398.

641 Chu H., Flerer N., Lauber C.L., Caporaso J.G., Knight R., Grogan P. (2010). Soil bacterial
642 diversity in the Arctic is not fundamentally different from that found in other biomes.
643 *Environment. Microbiol.* 12, 2998-3006.

644 Cole J.J., Pace M.L., Caraco N.F., Steinhart G.S. (1993). Bacterial biomass and cell size
645 distributions More and larger cells in anoxic waters in lakes. *Aquatic Microbial Ecology* 38,
646 1627–1632.

647 Crump B.C., Amaral-zettler L.A., Kling G.W. (2012). Microbial diversity in arctic
648 freshwaters is structured by inoculation of microbes from soils. *The ISME Journal* 6, 1629–
649 1639.

650 Daims H., Lebedeva E.V., Pjevac P., Han P., Herbold C., Albertsen M., Jehmlich N.,
651 Palatinszky M., Vierheilig J., Bulaev A., Kirkegaard R.H., von Bergen M., Rattei T.,
652 Bendinger B., Nielsen P.H., Wagner M. (2015). Complete nitrification by *Nitrospira* bacteria.
653 *Nature.* *Nature* 528, 504–509.

654 Dittmar T., Kattner G. (2003). The biogeochemistry of the river and shelf ecosystem of the
655 Arctic Ocean: a review. *Marine Chemistry* 83, 103–120.

656 Dolnicki P., Grabiec M., Puczko D., Gawor Ł., Budzik T., Klementowski J. (2013).
657 Variability of temperature and thickness of permafrost active layer at coastal sites of
658 Svalbard. *Polish Polar Research* 34, 353–374.

659 Edwards A., Irvine-Fynn T., Mitchell A.C., Rassner S.M.E. (2014). A germ theory for glacial
660 systems? *WIREs Water* 1, 331–340.

661 Elmendorf S.C. et al. (2012). Plot-scale evidence of tundra vegetation change and links to
662 recent summer warming. *Nature Climate Change* 2, 453–457.

663 Fagerbakke K.M., Heldal M., Norland S. (1996). Content of carbon, nitrogen, oxygen, sulfur
664 and phosphorus in native aquatic and cultured bacteria. *Aquatic Microbial Ecology* 10, 15–27.

665 Finneran K.T., Johnsen C.V., Lovley D.R. (2003). *Rhodoferrax ferrireducens* sp. nov., a
666 sychrotolerant, facultatively anaerobic bacterium that oxidizes acetate with the reduction of
667 Fe(III). *Int. J. Syst. Evol. Microbiol.* 53, 669–673.

668 Frey K.E., McClelland J.W. (2009). Impacts of permafrost degradation on arctic river
669 biogeochemistry. *Hydrological Processes* 23, 169–182.

670 Gardner A.S. et al. (2013). A reconciled estimate of glacier contributions to sea level rise:
671 2003 to 2009. *Science (New York, N.Y.)* 340, 852–7.

672 Gawor J., Grzesiak J., Sasin-Kurowska J., Borsuk P., Gromadka R., Górniak D., Świętecki
673 A., Aleksandrak-Piekarczyk T., Zdanowski M. K. (2016). Evidence of adaptation, niche
674 separation and microevolution within the genus *Polaromonas* on Arctic and Antarctic glacial
675 surfaces. *Extremophiles* 20, 403–413.

676 Göransson H., Venterink H.O., Bååth E. (2011). Soil bacterial growth and nutrient limitation
677 along a chronosequence from a glacier forefield. *Soil Biology and Biochemistry* 43, 1333–
678 1340.

- 679 Górnjak D., Marszałek H., Jankowska K., Dunalska J. (2016). Bacterial community
680 succession in an arctic lake-stream system (Bratteg valley, SW Spitsbergen). *Boreal*
681 *Environment Research* 21, 115–133.
- 682 Granéli W., Bertilsson S., Philibert A. (2004). Phosphorus limitation of bacterial growth in
683 high Arctic lakes and ponds. *Aquatic Sciences* 66, 430–439.
- 684 Grube M., Berg G. (2009). Microbial consortia of bacteria and fungi with focus on the lichen
685 symbiosis. *Fungal Biol. Rev.* 23, 72-85.
- 686 Hell K., Edwards A., Zarsky J., Podmirseg S., Girdwood S., Pachebat J., Insam H., Sattler B.
687 (2013). The dynamic bacterial communities of a melting High Arctic glacier snowpack. *ISME*
688 *J.* 7, 1814–26.
- 689 Hem J.D., Cropper W.H. (1962). Survey of ferrous-ferric chemical equilibria and redox
690 potentials. In *Chemistry of Iron in Natural Water* [Report US Geological Survey], Udall SL
691 and Nolan TB (eds). Washington, US; 1–31.
- 692 Hobbie J.E., Corliss T.L., Peterson B.J. (1983). Seasonal patterns of bacterial abundance in an
693 arctic lake. *Arctic and Alpine Research* 15, 253–260.
- 694 Hodson A.J., Mumford P.N., Kohler J., Wynn P.M. (2005). The High Arctic glacial
695 ecosystem: new insights from nutrient budgets. *Biogeochemistry* 72, 233–256.
- 696 Hood E., McKnight D.M., Williams M.W. (2003). Sources and chemical character of
697 dissolved organic carbon across an alpine/subalpine ecotone. Green Lakes Valley. Colorado
698 Front Range, United States. *Water Resour. Res.* 39, 1188.
- 699 Houghton J.T., Ding Y., Griggs D.J., Noguera M., van der Linden P.J., Dai X., Maskell K.,
700 Johnson C.A. (2001). *Climate Change 2001: The Scientific Basis. Contribution of Working*

701 Group I to the Third Assessment Report of the Intergovernmental Panel on Climate Change.
702 Cambridge University Press: Cambridge, UK.

703 Hultman J., Waldrop M.P., Mackelprang R., David M.M., McFarland J., Blazewicz S.J., et al.
704 (2015). Multi-omics of permafrost, active layer and thermokarst bog soil microbiomes. *Nature*
705 521, 208–212.

706 Hung H., Kallenborn R., Breivik K., Su Y., Brorström-Lundén E., Olafsdottir K., Thorlacius
707 J.M., Leppanen S., Bossi R., Skov H., Mano S., Patton G.W., Stern G., Sverko E., Fellin P.
708 (2010). Atmospheric monitoring of organic pollutants in the Arctic under the Arctic
709 Monitoring and Assessment Programme (AMAP): 1993-2006. *Sci. Total Environ.* 408, 2854-
710 2873.

711 Hutz A., Schubert K., Overmann J. (2011). *Thalassospira sp* isolated from the oligotrophic
712 eastern mediterranean sea exhibits chemotaxis toward inorganic phosphate during starvation.
713 *Appl. Environ. Microbiol.* 77, 4412–4421.

714 Ivanova A.A., Kulichevskaya I.S., Merkel A.Y., Toshchakov S.V., Dedysh S.N. (2016). High
715 Diversity of Planctomycetes in Soils of Two Lichen-Dominated Sub-Arctic Ecosystems of
716 Northwestern Siberia. *Front Microbiol.* 22, 7-2065.

717 Jones S.E., Newton R.J., McMahon K.D. (2009). Evidence for structuring of bacterial
718 community composition by organic carbon source in temperate lakes. *Environ. Microbiol.*
719 11, 2463-2472.

720 Jørgensen L., Markager S., Maar M. (2014). On the importance of quantifying bioavailable
721 nitrogen instead of total nitrogen. *Biogeochemistry* 117, 455-472.



- 722 Kim H.M., Jung J.Y., Yergeau E., Hwang C.Y., Hinzman L., Nam S., et al. (2014). Bacterial
723 community structure and soil properties of a subarctic tundra soil in Council, Alaska. *FEMS*
724 *Microbiol. Ecol.* 89, 465–475.
- 725 Kirsten G., Brett K.S., Ari J., Lydia H.Z. (2018). The Avian Gut Microbiota: Community,
726 Physiology and Function in Wild Birds. *Journals of Avian Biology*. First published: 26 July
727 2018. doi: [10.1111/jav.01788].
- 728 Klindworth A., Pruesse E., Schweer T., Peplies J., Quast C., Horn M., Glöckner F.O. (2012).
729 Evaluation of general 16S ribosomal RNA gene PCR primers for classical and next-
730 generation sequencing-based diversity studies. *Nucleic Acids Res.* 7, 41(1).
- 731 Kolton M., Sela N., Elad Y., Cytryn E. (2013). Comparative genomic analysis indicates that
732 niche adaptation of terrestrial Flavobacteria is strongly linked to plant glycan metabolism.
733 *PLoS One*. Sep 26, 8(9):e76704. doi: 10.1371/journal.pone.0076704. eCollection 2013.
- 734 Kolton M., Erlacher A., Berg G., Cytryn E. (2016). The *Flavobacterium* Genus in the Plant
735 Holobiont: Ecological, Physiological, and Applicative Insights. *Microbial Models: From*
736 *Environmental to Industrial Sustainability* 1, 189-207.
- 737 Kosek K., Jankowska K., Polkowska Ż. (2017). Bacterial presence in polar regions associated
738 with environment modification by chemical compounds including contaminants. *Environ.*
739 *Rev.* 25, 481-491.
- 740 Kosek K., Kozak K., Koziół K., Jankowska K., Chmiel S., Polkowska Ż. (2018). The
741 interaction between bacterial abundance and selected pollutants concentration levels in an
742 Arctic catchment (southwest Spitsbergen, Svalbard). *Sci. Total Environ.* 622-623, 913-923.

743 Kosek K., Koziół K., Luczkiewicz A., Jankowska K., Chmiel S., Polkowska Z. (accepted).
744 Abundance and bacterial community shaped by the environmental characteristics of an Arctic
745 tundra river system (southwestern Spitsbergen, Svalbard). Part 2: Chemical stress factors.

746 Kozak K., Koziół K., Luks B., Chmiel S., Ruman M., Marć M., Namieśnik J., Polkowska Ż.
747 (2015). The role of atmospheric precipitation in introducing contaminants to the surface
748 waters of the Fuglebekken catchment, Spitsbergen. *Polar Research* 34, 24207.

749 Kozak K., Polkowska Ż., Stachnik Ł., Luks B., Chmiel S., Ruman M., Lech D., Koziół K.,
750 Tsakovski S., Simeonov V. (2016). Arctic catchment as a sensitive indicator of the
751 environmental changes: distribution and migration of metals (Svalbard). *Int. J. Environ. Sci.*
752 *Technol.* 13, 2779-2796.

753 Kühnel R., Björkman M.P., Vega C.P., Hodson A.J., Isaksson E., Ström J. (2013). Reactive
754 nitrogen and sulphate wet deposition at Zeppelin Station, Ny-Ålesund, Svalbard. *Polar*
755 *Research* 32, 19136.

756 Lee K., Kim T-W., Byrne R.H., Millero F.J., Feely R.A., Liu Y.M. (2010). The universal ratio
757 of boron to chlorinity for the North Pacific and North Atlantic oceans. *Geochimica et*
758 *Cosmochimica Acta* 74, 1801–1811.

759 Lotti T., Kleerebezem R., van Erp Taalman Kip C., Hendrickx T.L., Kruit J., Hoekstra M.,
760 van Loosdrecht M.C. (2014). Anammox growth on pretreated municipal wastewater. *Environ*
761 *Sci Technol.* 48, 7874-7880.

762 Łupikasza E. (2007). Long-term variability of extreme precipitation in Hornsund
763 (Spitsbergen) and their relation with atmospheric circulation. *Problemy Klimatologii Polarnej*
764 17, 87–103.

765 Männistö M.K., Kurhela E., Tiirola M., Häggblom M.M. (2013). Acidobacteria dominate the
766 active bacterial communities of Arctic tundra with widely divergent winter-time snow
767 accumulation and soil temperatures, *FEMS Microbiology Ecology*, Volume 84, 1, 47–59.

768 Marszałek H., Wąsik M. (2013). Some physico-chemical features of water in suprapermafrost
769 zone in the Hornsund region (SW Spitsbergen). *Biuletyn Państwowego Instytutu*
770 *Geologicznego* 456, 397-404.

771 McKenzie J.M., Voss C.I. (2013). Permafrost thaw in a nested groundwater-flow system.
772 *Hydrogeology Journal* 21, 299-316.

773 McNamara J.P., Kane D.L., Hobbie J.E., Kling G.W. (2008). Hydrologic and biogeochemical
774 controls on the spatial and temporal patterns of nitrogen and phosphorus in the Kuparuk
775 River, arctic Alaska. *Hydrological Processes* 22, 3294–3309.

776 Mindl B., Anesio A.M., Meirer K., Hodson A.J., Laybourn-Parry J., Sommaruga R., Sattler B.
777 (2007). Factors influencing bacterial dynamics along a transect from supraglacial runoff to
778 proglacial lakes of a high Arctic glacier [corrected]. *FEMS Microbiology Ecology* 59, 307–
779 17.

780 Neufeld J.D., Mohn W.W. (2005). Unexpectedly high bacterial diversity in arctic tundra
781 relative to boreal forest soils, revealed by serial analysis of ribosomal sequence tags. *Appl.*
782 *Environ. Microbiol.* 71, 5710-5718.

783 Nissinen R.M., Männistö M.K., van Elsas J.D. (2012). Endophytic bacterial communities in
784 three arctic plants from low arctic fell tundra are cold-adapted and host-plant specific. *FEMS*
785 *Microbiol Ecol.* 82, 510-22.

786 Nixon S.L., Telling J., Wadham J.L., Cockell C.S., Nixon S. (2017). Viable cold-tolerant
787 iron-reducing microorganisms in geographically-isolated subglacial environments.
788 *Biogeosciences Discuss.* 14, 1445-1455.

789 Ntougias S., Polkowska Ż., Nikolaki S., Dionyssopoulou E., Stathopoulou P., Doudoumis V.,
790 Ruman M., Kozak K., Namieśnik J., Tsiamis G. (2016). Bacterial Community Structures in
791 Freshwater Polar Environments of Svalbard. *Microbes. Environ* 31, 401-409.

792 Petrone K.C., Richards J.S. (2009). Bioavailability and composition of dissolved organic
793 carbon and nitrogen in a near coastal catchment of south-western Australia. *Biogeochemistry*
794 92, 27-40.

795 Pomeroy L.R., Wiebe W.J. (2001). Temperature and substrates as interactive limiting factors
796 for marine heterotrophic bacteria. *Aquat. Microb. Ecol.* 23, 187-204.

797 Pulina M., Krawczyk W., Pereyma J. (1984). Water balance and chemical denudation in the
798 unglaciated Fugleberget basin (SW Spitsbergen). *Polish Polar Research* 5, 183–205.

799 Rawat S.R., Männistö M.K., Bromberg Y., Häggblom M.M. (2012). Comparative genomic
800 and physiological analysis provides insights into the role of Acidobacteria in organic carbon
801 utilization in Arctic tundra soils. *FEMS Microbiol Ecol.* 82, 2, 341-55.

802 Ruman M., Kozak K., Lehmann S., Koziol K., Polkowska Ż. (2012). Pollutants present In
803 different components of the Svalbard Archipelago environment. *Ecol. Chem. Eng. S.* 19, 571-
804 584.

805 Šimek K., Vrba J., Hartman P. (1994). Size-Selective Feeding by *Cyclidium* sp. on
806 Bacterioplankton and Various Sizes of Cultured Bacteria. *FEMS Microbiology Ecology* 14,
807 157–167.

808 Smulikowski W. (1965). Petrology and some structural data of lower metamorphic formations
809 of the Hecla Hoek Succession in Hornsund, Vestspitsbergen, *Studia Geologica Polonica*,
810 Warsaw 18, 1-107.

811 Stibal M., Tranter M., Telling J., Benning L.G. (2008). Speciation, phase association and
812 potential bioavailability of phosphorus on Svalbard glacier. *Biogeochemistry* 90, 1-13.

813 Szumińska D., Szopińska M., Lehmann-Konera S., Franczak Ł., Kociuba W., Chmiel S.,
814 Kalinowski P., Polkowska Ż. (2018). Water chemistry of tundra lakes in the periglacial zone
815 of the Bellsund Fiord (Svalbard) in the summer 2013. *Sci. Total. Environ.* 624, 1669-1679.

816 Tveit A., Schwacke R., Svenning M.M., Urich T. (2013). Organic carbon transformation in
817 high-Arctic peat soils: key functions and microorganisms. *ISME J.* 7, 299-311.

818 Vallières C., Retamal L., Ramlal P., Osburn C.L., Vincent W.F. (2008). Bacterial production
819 and microbial food web structure in a large arctic river and the coastal Arctic Ocean. *Journal*
820 *of Marine Systems* 74, 756–773.

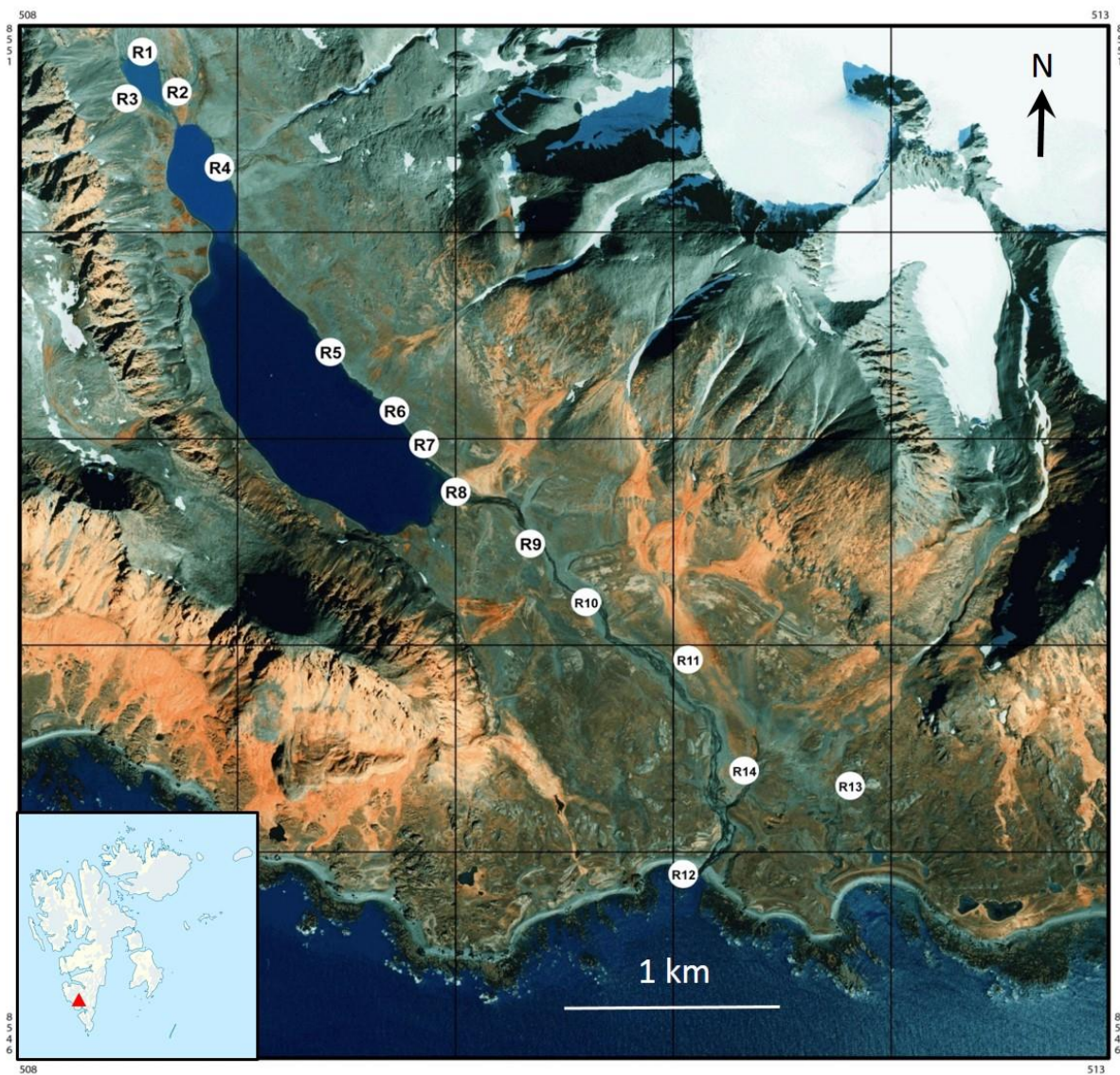
821 Vrede K., Heldal M., Norland S., Bratbak G. (2002). Elemental Composition (C, N, P) and
822 Cell Volume of Exponentially Growing and Elemental Composition (C, N, P) and Cell
823 Volume of Exponentially Growing and Nutrient-Limited Bacterioplankton. *Applied and*
824 *Environmental Microbiology* 68, 2965-2971.

825 Wagner A., Grillitsch K., Leitner E., Daum G. (2009). Mobilization of steryl esters from lipid
826 particles of the yeast *Sacchromyces cerevisiae*. *Biochem. Biophys. Acta* 1791, 118-124.

827 Wallenstein M., McMahon S.K., Schimel J.P. (2009). Seasonal Variation in Enzyme
828 Activities and Temperature Sensitivities in Arctic Tundra Soils. *Global Change Biology* 15,
829 1631-1639.

830 Wojciechowski J. (1964). Ore-bearing veins of the Hornsund area, Vestspitsbergen. *Studia*
831 *Geol. Pol.* 11, 173-177.

832

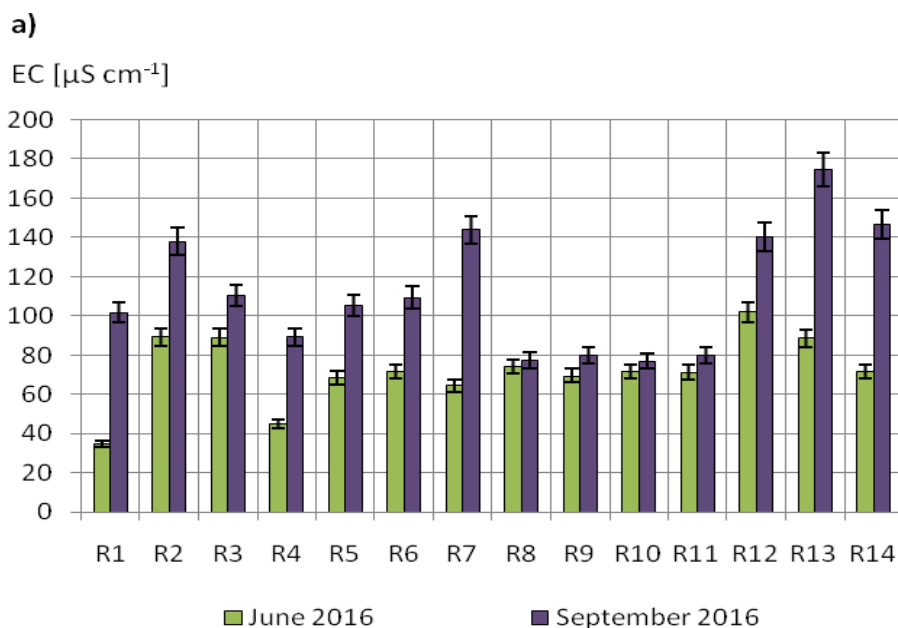


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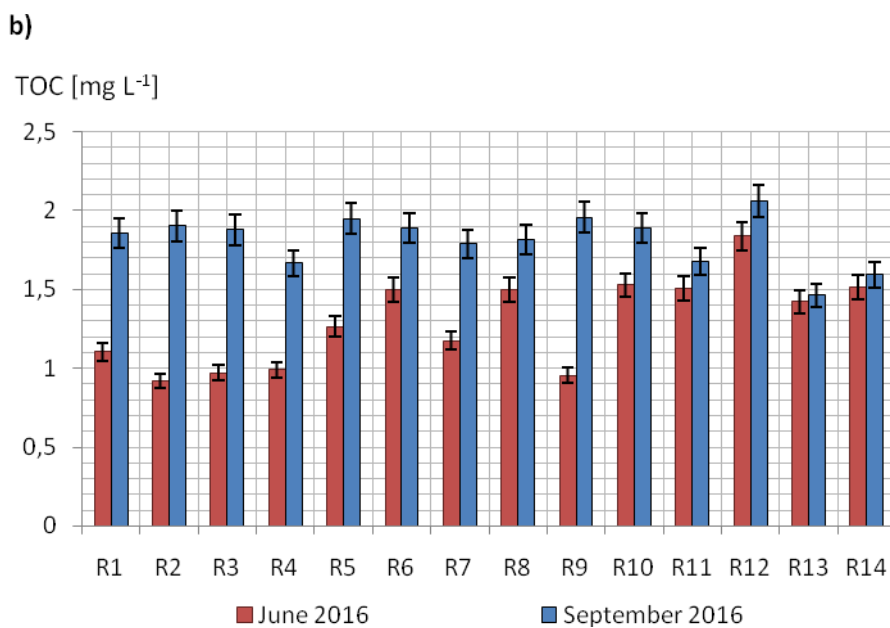
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835 **Figure 1.** Location of the studied area in Svalbard and the sampling points in the Revelva

836 catchment.



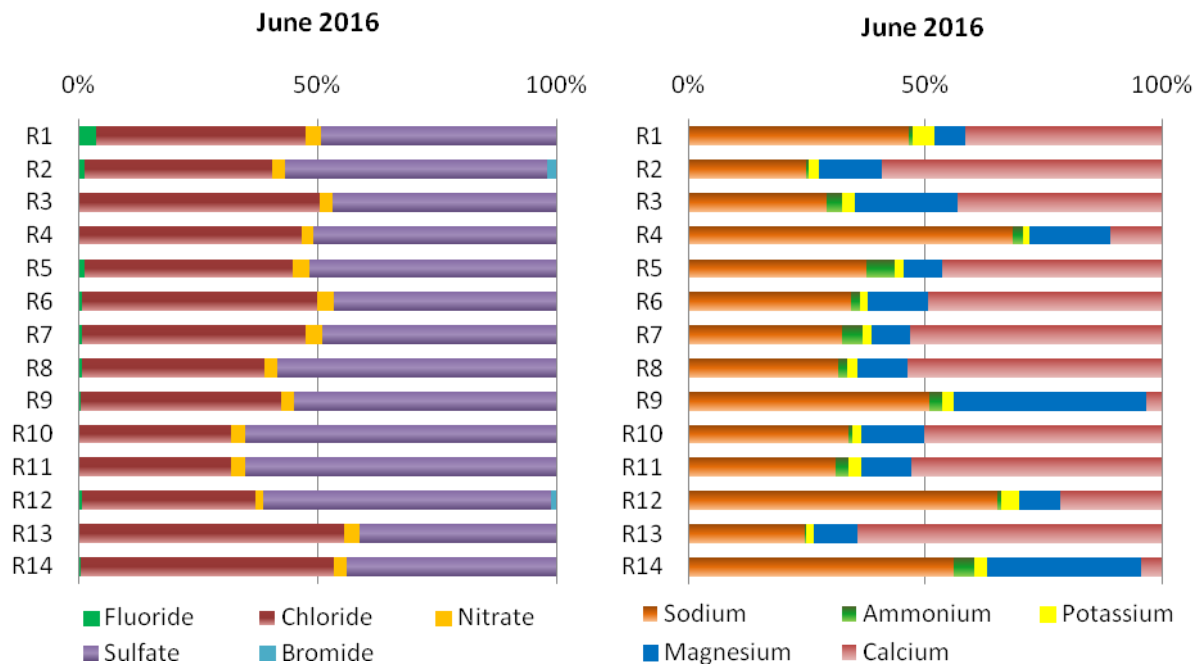
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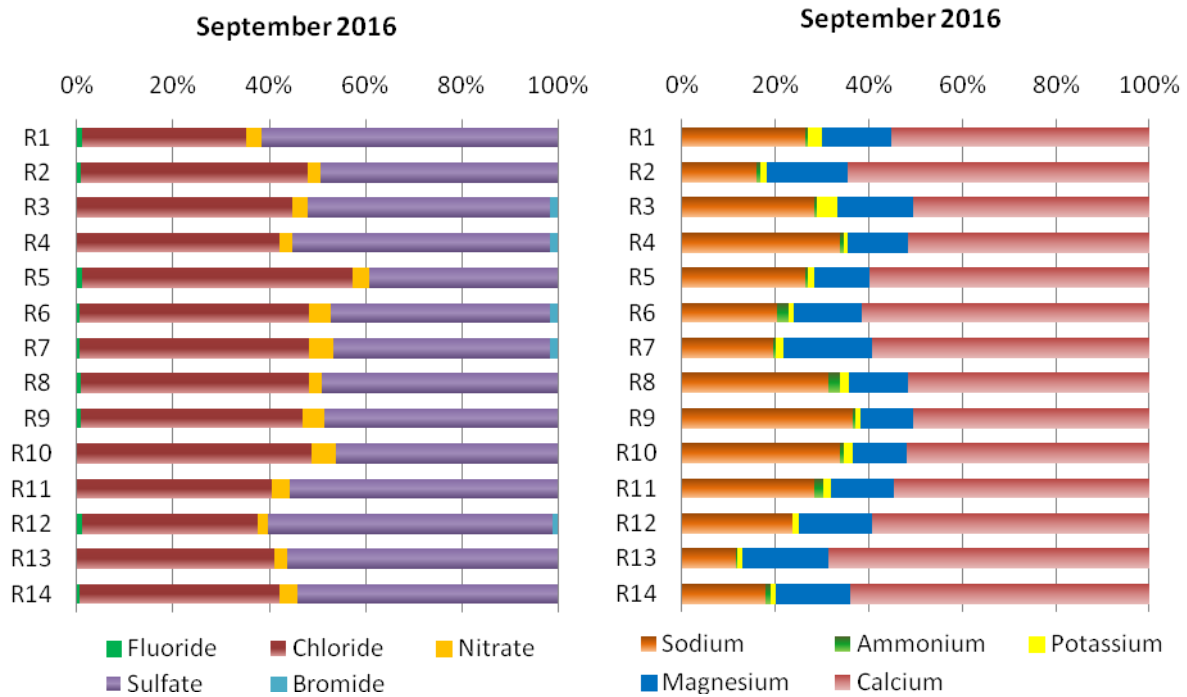
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840 **Figure 2.** Concentration levels of electrical conductivity and total organic carbon determined
 841 in the collected freshwater samples, compared between the studied periods; a) electrical
 842 conductivity (EC), b) total organic carbon (TOC).

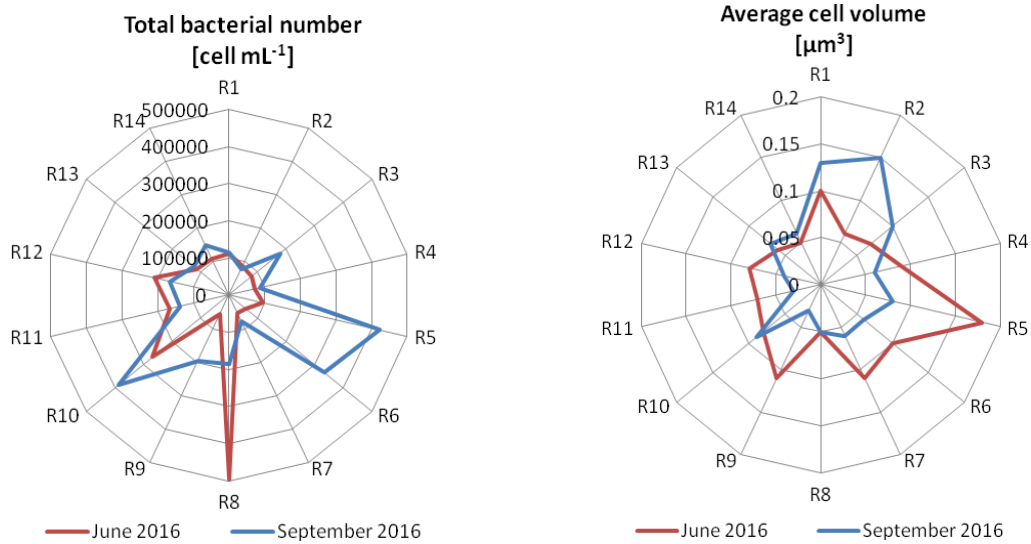


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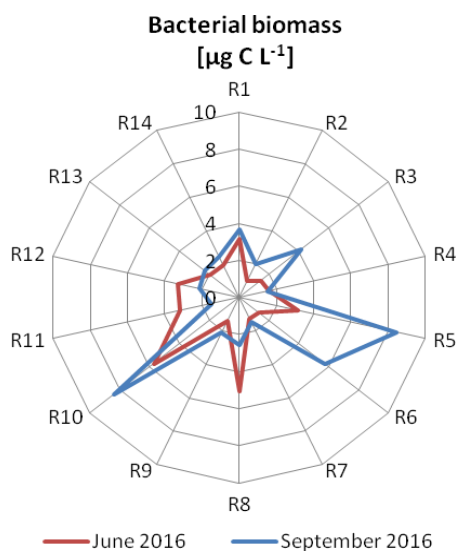


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845 **Figure 3.** Percentage anion and cation composition of the collected samples.

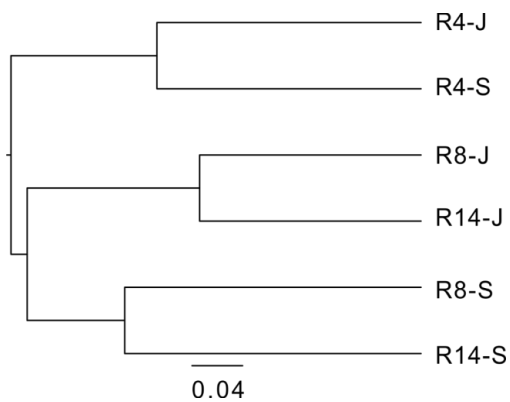


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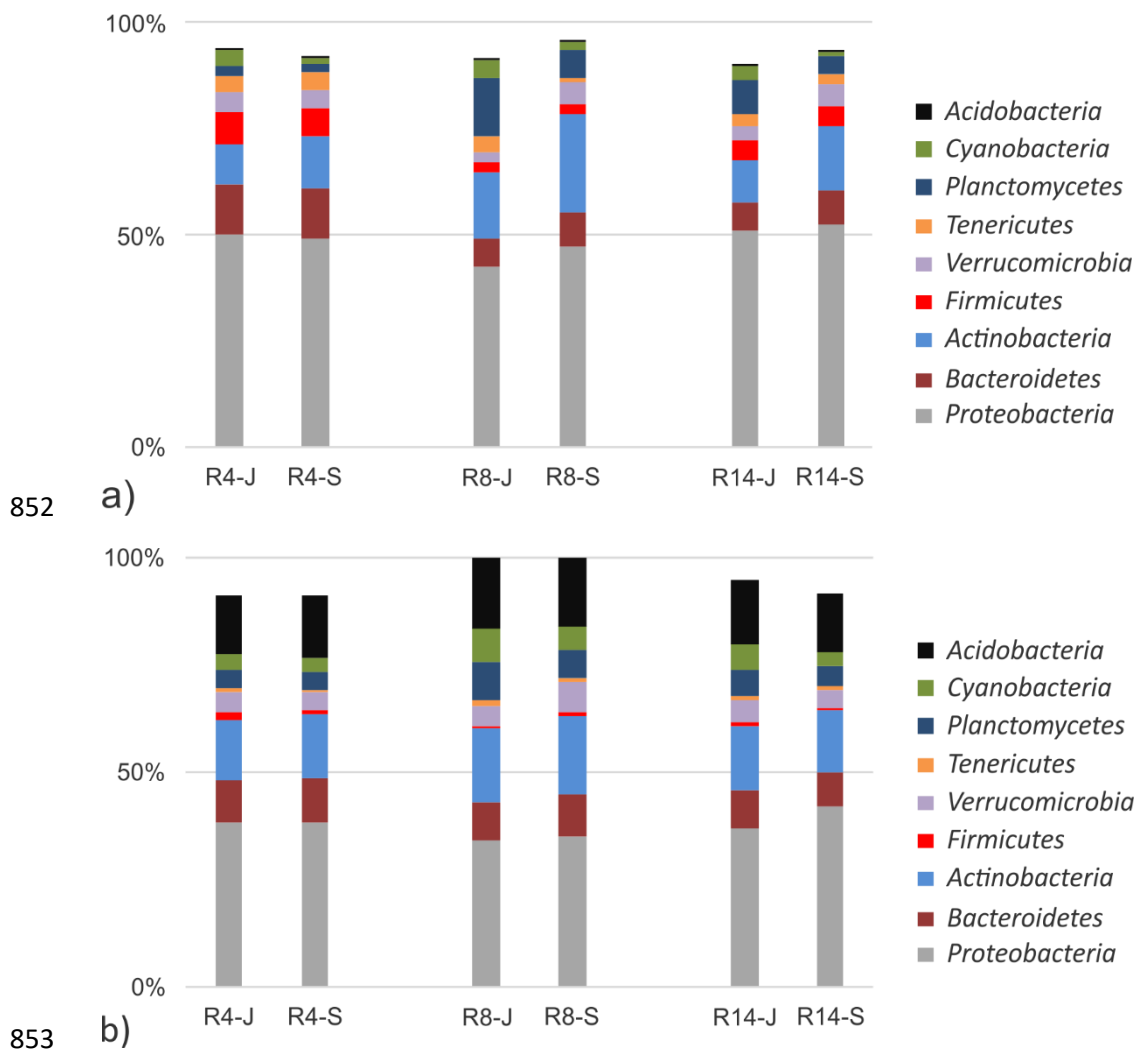
847

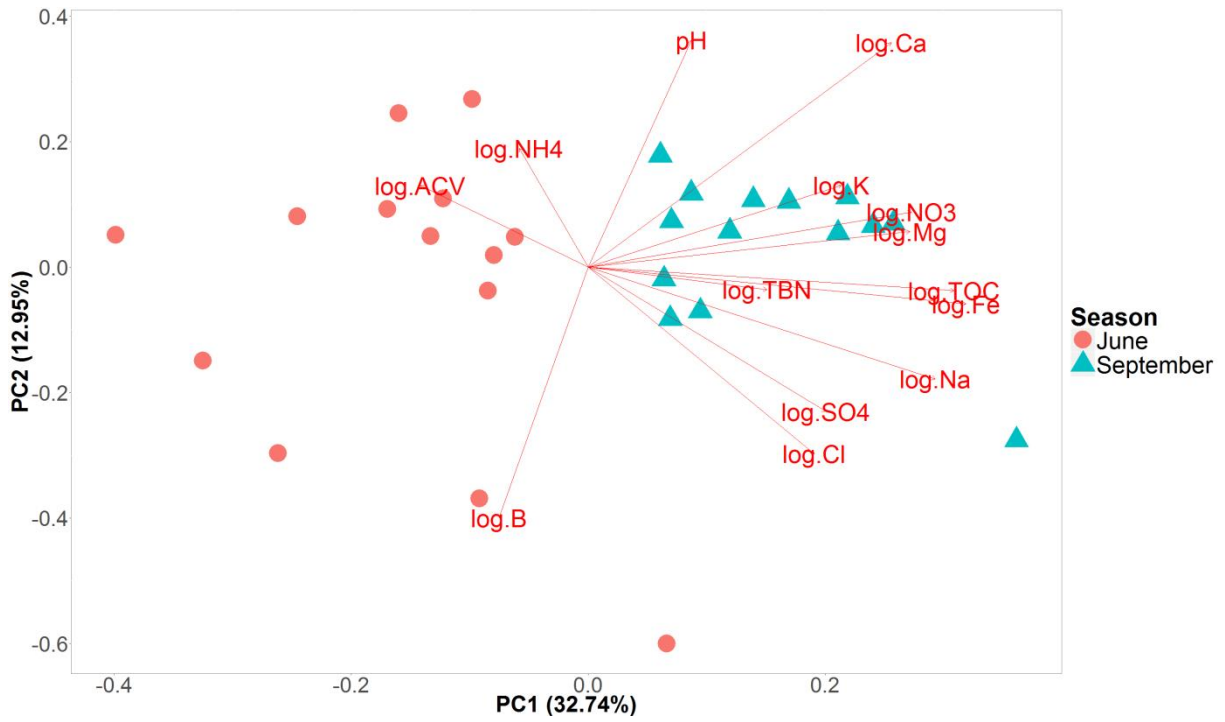
848 **Figure 4.** Comparison of bacterial abundance (total number), average bacterial cell volume
 849 and bacterial biomass in the Revelva catchment in June and September 2016.



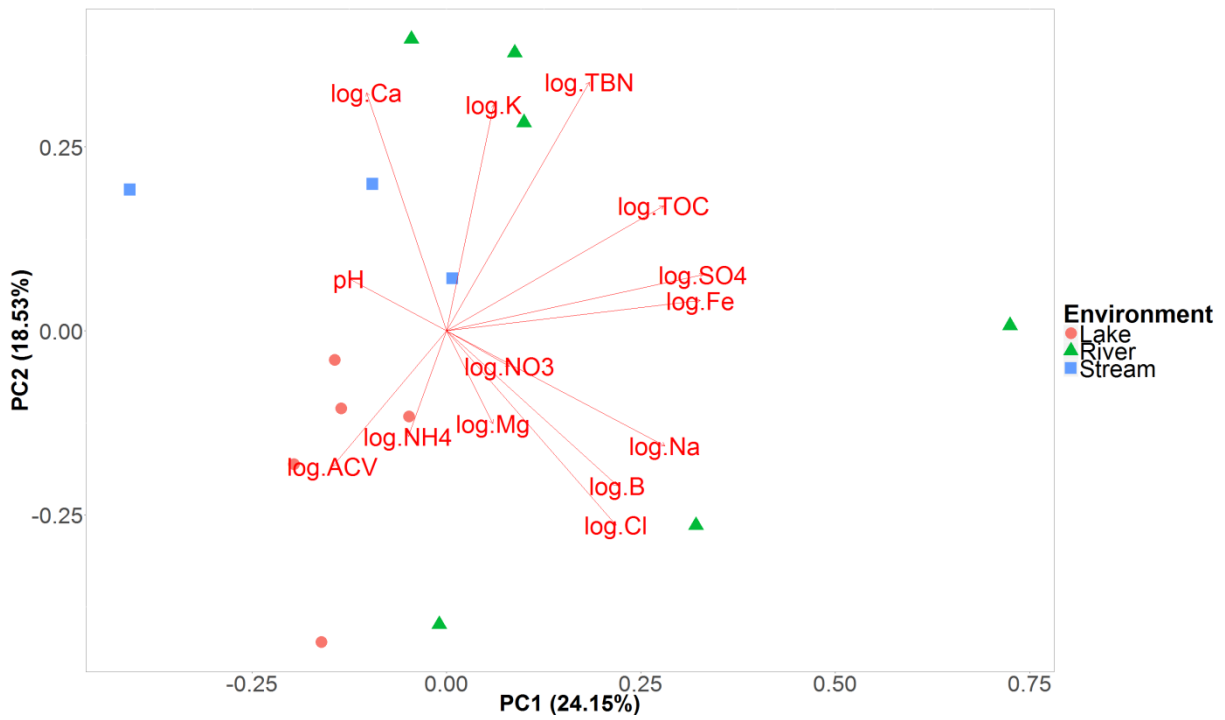
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851 **Figure 5.** Cluster analysis of bacterial community structures.

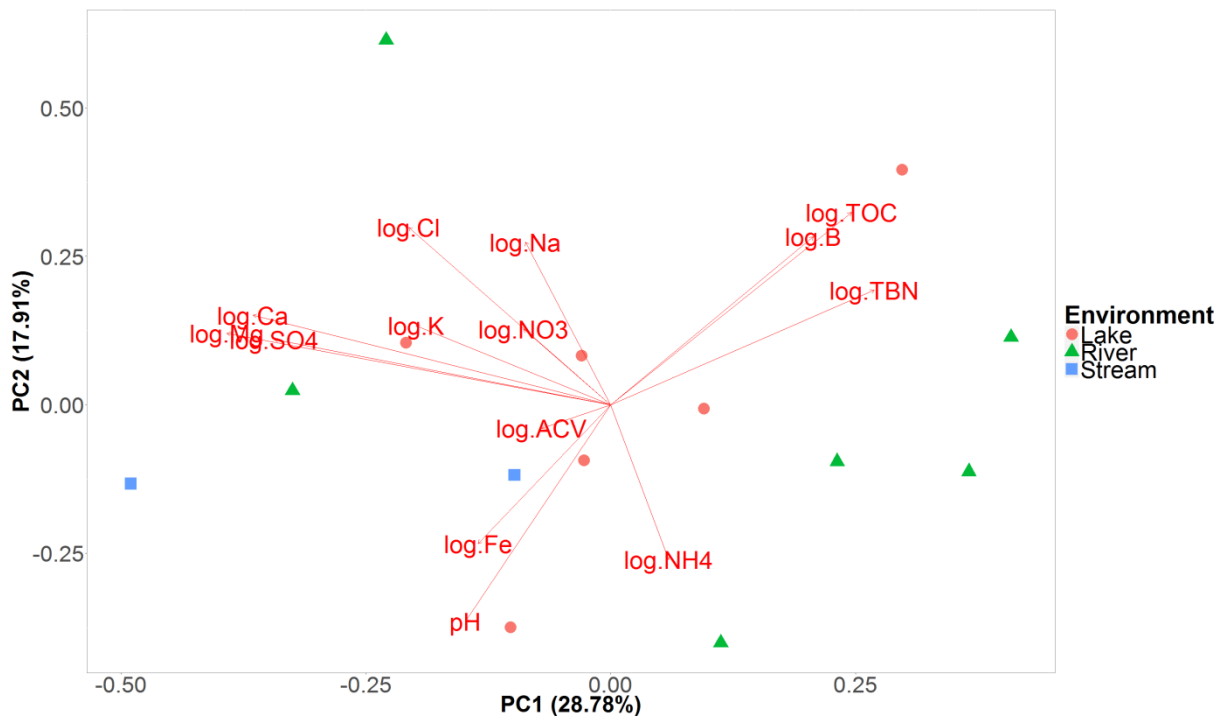




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860 **Figure 7.** Principal component analysis results for nutrient concentrations and the bacterial
 861 community parameters (top). The two graphs below represent the two studied periods (June in
 862 the middle and September at the bottom), which were clearly divided in the analysis of the
 863 whole dataset.

864 **Tables**865 **Table 1.** Validation parameters and technical specifications used in the applied analytical procedures.

Determined compounds/parameters	Measurement range	LOD ⁴	LOQ ⁴	Measurement method/technique
Electrical conductivity ¹	-	-	-	Electrochemical method: CPC-411 conductometer (Elmetron), conductivity sensor EC60
pH	-	-	-	Electrochemical method: microcomputer pH-meter(Elmetron), electrode type EPS-1
TOC ²	0.150-10.0	0.030	0.100	Total Organic Carbon Analyzer, TOC-V _{CSH/CSN} , method of catalytic combustion (oxidation) with the application of the NDIR detector
Anions ²	0.030-250	0.060	0.180	Ion Chromatography technique with the application of the conductivity detector (DIONEX ICS-3000)
Cations ²	0.030-250	0.010	0.030	
Micronutrients ³	Fe	0.010-1000	0.010	Inductively Coupled Plasma Mass Spectrometry technique (Thermo Scientific XSERIES 2 ICP-MS)
	B	0.100-1000	0.100	
	P	1.00-1000	1.00	

866 ¹[$\mu\text{S cm}^{-1}$], ²[mg L^{-1}], ³[$\mu\text{g L}^{-1}$], ⁴the limit of detection (LOD) and the limit of quantification (LOQ) were calculated based on the standard deviation of the response (s) and the867 slope of the calibration curve (b), according to the formulas: $\text{LOD}=3.3(\text{s}/\text{b})$, $\text{LOQ}=10(\text{s}/\text{b})$ 

868 **Table 2.** Concentrations (\pm standard deviation, SD) of micronutrients in the collected freshwater samples.

		June 2016	September 2016
Elemental nutrients	B	1.701 \pm 0.041 – 4.513 \pm 0.024	1.888 \pm 0.038 – 2.714 \pm 0.016
[$\mu\text{g L}^{-1}$]	P	<LOD – 1.91 \pm 0.47	<LOD
	Fe	0.0100 \pm 0.0050 – 0.279 \pm 0.017	0.1210 \pm 0.0010 – 1.32 \pm 0.14

869 **Table 3.** Number of reads and OTUs as well as species richness estimate (Chao1) and diversity indices (Shannon and Simpson) for the sampling
870 points.

Sampling point	Reads	OTU	Chao 1	Shannon	Simpson
R4-J	154022	6500	6755	10.32	0.995
R4-S	170636	6732	7026	10.42	0.996
R8-J	118556	3838	4914	7.28	0.971
R8-S	119228	3936	5485	6.82	0.948
R14-J	99569	5576	6284	9.13	0.987
R14-S	178337	6663	6920	9.89	0.992

871 OTUs were defined at a 97% sequence identity threshold

872

873 **Table 4.** Family level taxonomic composition in the Revelva catchment (among 274 families reported in this study, first 40 are presented).

R4 in June			R4 in September			R8 in June			R8 in September			R14 in June			R14 in September		
Family	%	%	Family	%	%	Family	%	%	Family	%	%	Family	%	%	Family	%	%
<i>Comamonadaceae</i>	11.66%	13.00%	<i>Comamonadaceae</i>			<i>Isosphaeraceae</i>	11.53%	13.45%	<i>Kiloniellaceae</i>			<i>Comamonadaceae</i>	6.48%	8.19%	<i>Comamonadaceae</i>		
<i>Flavobacteriaceae</i>	5.57%	3.59%	<i>Flavobacteriaceae</i>			<i>Streptomyetaceae</i>	7.68%	11.21%	<i>Cellulomonadaceae</i>			<i>Isosphaeraceae</i>	6.35%	5.22%	<i>Cellulomonadaceae</i>		
<i>Oxalobacteraceae</i>	4.46%	3.16%	<i>Flexibacteraceae</i>			<i>Comamonadaceae</i>	6.35%	10.64%	<i>Comamonadaceae</i>			<i>Oxalobacteraceae</i>	5.30%	3.79%	<i>Kiloniellaceae</i>		
<i>Clostridiaceae</i>	2.48%	2.67%	<i>Sphingomonadaceae</i>			<i>Oxalobacteraceae</i>	5.64%	5.99%	<i>Isosphaeraceae</i>			<i>Kiloniellaceae</i>	4.76%	3.43%	<i>Oxalobacteraceae</i>		
<i>Flexibacteraceae</i>	2.22%	2.56%	<i>Intrasporangiaceae</i>			<i>Kiloniellaceae</i>	5.31%	4.16%	<i>Verrucomicrobiaceae</i>			<i>Flavobacteriaceae</i>	3.74%	3.12%	<i>Verrucomicrobiaceae</i>		
<i>Sphingobacteriaceae</i>	1.97%	2.54%	<i>Sphingobacteriaceae</i>			<i>Flavobacteriaceae</i>	4.73%	3.86%	<i>Oxalobacteraceae</i>			<i>Cellulomonadaceae</i>	2.68%	2.92%	<i>Isosphaeraceae</i>		
<i>Sphingomonadaceae</i>	1.90%	2.15%	<i>Acholeplasmataceae</i>			<i>Rhizobiaceae</i>	2.93%	3.52%	<i>Pseudonocardiaceae</i>			<i>Legionellaceae</i>	1.88%	2.64%	<i>Legionellaceae</i>		
<i>Geobacteraceae</i>	1.74%	2.12%	<i>Xanthomonadaceae</i>			<i>Cellulomonadaceae</i>	2.01%	2.85%	<i>Flavobacteriaceae</i>			<i>Pseudonocardiaceae</i>	1.86%	2.39%	<i>Intrasporangiaceae</i>		
<i>Bifidobacteriaceae</i>	1.72%	1.97%	<i>Cellulomonadaceae</i>			<i>Mycoplasmataceae</i>	1.93%	2.76%	<i>Chitinophagaceae</i>			<i>Rhizobiaceae</i>	1.78%	2.24%	<i>Flavobacteriaceae</i>		
<i>Xanthomonadaceae</i>	1.69%	1.83%	<i>Oxalobacteraceae</i>			<i>Pseudonocardiaceae</i>	1.76%	1.90%	<i>Microbacteriaceae</i>			<i>Sphingomonadaceae</i>	1.33%	2.23%	<i>Rhodospirillaceae</i>		
<i>Legionellaceae</i>	1.53%	1.67%	<i>Pseudonocardiaceae</i>			<i>Brocadiaceae</i>	1.58%	1.42%	<i>Flexibacteraceae</i>			<i>Mycoplasmataceae</i>	1.26%	2.10%	<i>Sphingomonadaceae</i>		
<i>Intrasporangiaceae</i>	1.49%	1.66%	<i>Verrucomicrobiaceae</i>			<i>Verrucomicrobiaceae</i>	1.52%	1.13%	<i>Legionellaceae</i>			<i>Brocadiaceae</i>	1.21%	2.04%	<i>Pseudonocardiaceae</i>		
<i>Chitinophagaceae</i>	1.37%	1.61%	<i>Chitinophagaceae</i>			<i>Legionellaceae</i>	1.37%	1.10%	<i>Veillonellaceae</i>			<i>Verrucomicrobiaceae</i>	1.10%	1.80%	<i>Sphingobacteriaceae</i>		
<i>Rhodospirillaceae</i>	1.27%	1.49%	<i>Rhodocyclaceae</i>			<i>Rivulariaceae</i>	1.18%	1.00%	<i>Methylophilaceae</i>			<i>Hyphomicrobiaceae</i>	1.08%	1.75%	<i>Acholeplasmataceae</i>		
<i>Hyphomicrobiaceae</i>	1.10%	1.40%	<i>Rhodospirillaceae</i>			<i>Halothiobacillaceae</i>	1.01%	0.99%	<i>Alcaligenaceae</i>			<i>Rhodospirillaceae</i>	1.05%	1.68%	<i>Chitinophagaceae</i>		
<i>Thermoanaerobacteraceae</i>	1.03%	1.40%	<i>Paenibacillaceae</i>			<i>Alcaligenaceae</i>	0.68%	0.97%	<i>Bogoriellaceae</i>			<i>Geobacteraceae</i>	0.97%	1.54%	<i>Flexibacteraceae</i>		
<i>Pseudonocardiaceae</i>	1.01%	1.32%	<i>Legionellaceae</i>			<i>Enterobacteriaceae</i>	0.66%	0.89%	<i>Sphingomonadaceae</i>			<i>Rhodocyclaceae</i>	0.94%	1.26%	<i>Xanthomonadaceae</i>		
<i>Puniceicoccaceae</i>	1.01%	1.23%	<i>Clostridiaceae</i>			<i>Cyanobacteriaceae</i>	0.65%	0.64%	<i>Caulobacteraceae</i>			<i>Flexibacteraceae</i>	0.92%	1.18%	<i>Caulobacteraceae</i>		
<i>Coxiellaceae</i>	1.00%	1.21%	<i>Hyphomicrobiaceae</i>			<i>Micromonosporaceae</i>	0.62%	0.64%	<i>Rhodobacteraceae</i>			<i>Coxiellaceae</i>	0.90%	1.15%	<i>Hyphomicrobiaceae</i>		
<i>Rhodocyclaceae</i>	0.99%	1.09%	<i>Thermoanaerobacteraceae</i>			<i>Geobacteraceae</i>	0.57%	0.50%	<i>Rhodocyclaceae</i>			<i>Alcaligenaceae</i>	0.89%	1.05%	<i>Microbacteriaceae</i>		
<i>Verrucomicrobiaceae</i>	0.99%	0.97%	<i>Microbacteriaceae</i>			<i>Caulobacteraceae</i>	0.55%	0.50%	<i>Streptomyetaceae</i>			<i>Sphingobacteriaceae</i>	0.86%	0.97%	<i>Rhodocyclaceae</i>		
<i>Cellulomonadaceae</i>	0.92%	0.96%	<i>Geobacteraceae</i>			<i>Flexibacteraceae</i>	0.55%	0.49%	<i>Intrasporangiaceae</i>			<i>Chitinophagaceae</i>	0.84%	0.92%	<i>Geobacteraceae</i>		
<i>Brocadiaceae</i>	0.91%	0.85%	<i>Chromatiaceae</i>			<i>Planctomycetaceae</i>	0.50%	0.43%	<i>Hyphomonadaceae</i>			<i>Intrasporangiaceae</i>	0.78%	0.79%	<i>Coxiellaceae</i>		
<i>Paenibacillaceae</i>	0.89%	0.84%	<i>Caulobacteraceae</i>			<i>Coxiellaceae</i>	0.50%	0.42%	<i>Cyanobacteriaceae</i>			<i>Caulobacteraceae</i>	0.77%	0.75%	<i>Paenibacillaceae</i>		
<i>Chromatiaceae</i>	0.88%	0.78%	<i>Puniceicoccaceae</i>			<i>Sphingobacteriaceae</i>	0.47%	0.40%	<i>Polyangiaceae</i>			<i>Rivulariaceae</i>	0.76%	0.70%	<i>Chromatiaceae</i>		
<i>Chthoniobacteraceae</i>	0.84%	0.75%	<i>Brocadiaceae</i>			<i>Sphingomonadaceae</i>	0.45%	0.37%	<i>Thermogemmatimonadaceae</i>			<i>Enterobacteriaceae</i>	0.74%	0.67%	<i>Clostridiaceae</i>		
<i>Acholeplasmataceae</i>	0.80%	0.74%	<i>Methylophilaceae</i>			<i>Hyphomicrobiaceae</i>	0.45%	0.36%	<i>Armatimonadaceae</i>			<i>Xanthomonadaceae</i>	0.71%	0.66%	<i>Brocadiaceae</i>		
<i>Acetobacteraceae</i>	0.77%	0.63%	<i>Chthoniobacteraceae</i>			<i>Chthoniobacteraceae</i>	0.42%	0.34%	<i>Rhodospirillaceae</i>			<i>Puniceicoccaceae</i>	0.68%	0.65%	<i>Alcaligenaceae</i>		
<i>Caulobacteraceae</i>	0.75%	0.57%	<i>Bifidobacteriaceae</i>			<i>Methylophilaceae</i>	0.42%	0.33%	<i>Xanthomonadaceae</i>			<i>Paenibacillaceae</i>	0.68%	0.63%	<i>Enterobacteriaceae</i>		
<i>Microbacteriaceae</i>	0.74%	0.55%	<i>Gemmatimonadaceae</i>			<i>Phormidiaceae</i>	0.41%	0.32%	<i>Bacteriovoracaceae</i>			<i>Chromatiaceae</i>	0.65%	0.61%	<i>Mycoplasmataceae</i>		
<i>Peptococcaceae</i>	0.67%	0.50%	<i>Pedosphaeraceae</i>			<i>Pasteurellaceae</i>	0.41%	0.31%	<i>Sphingobacteriaceae</i>			<i>Clostridiaceae</i>	0.65%	0.61%	<i>Bifidobacteriaceae</i>		

<i>Alcaligenaceae</i>	0.65%	0.48%	<i>Leuconostocaceae</i>	<i>Chitinophagaceae</i>	0.38%	0.30%	<i>Acetobacteraceae</i>	<i>Chthoniobacteraceae</i>	0.63%	0.58%	<i>Thermoanaerobacteraceae</i>
<i>Opitutaceae</i>	0.61%	0.48%	<i>Coxiellaceae</i>	<i>Lactobacillaceae</i>	0.37%	0.28%	<i>Brocadiaceae</i>	<i>Microbacteriaceae</i>	0.56%	0.56%	<i>Veillonellaceae</i>
<i>Polyangiaceae</i>	0.59%	0.46%	<i>Acetobacteraceae</i>	<i>Rhodocyclaceae</i>	0.36%	0.28%	<i>Chthoniobacteraceae</i>	<i>Veillonellaceae</i>	0.56%	0.56%	<i>Puniceicoccaceae</i>
<i>Mycoplasmataceae</i>	0.56%	0.45%	<i>Polyangiaceae</i>	<i>Nostocaceae</i>	0.35%	0.27%	<i>Micromonosporaceae</i>	<i>Thermoanaerobacteraceae</i>	0.54%	0.49%	<i>Bacteriovoracaceae</i>
<i>Veillonellaceae</i>	0.55%	0.44%	<i>Kiloniellaceae</i>	<i>Veillonellaceae</i>	0.34%	0.27%	<i>Acholeplasmataceae</i>	<i>Micromonosporaceae</i>	0.50%	0.49%	<i>Rhodobacteraceae</i>
<i>Conexibacteraceae</i>	0.53%	0.44%	<i>Isosphaeraceae</i>	<i>Microbacteriaceae</i>	0.33%	0.26%	<i>Hyphomicrobiaceae</i>	<i>Methylophilaceae</i>	0.48%	0.47%	<i>Leuconostocaceae</i>
<i>Enterobacteriaceae</i>	0.52%	0.42%	<i>Opitutaceae</i>	<i>Mycobacteriaceae</i>	0.32%	0.26%	<i>Cerasicoccaceae</i>	<i>Acholeplasmataceae</i>	0.47%	0.47%	<i>Chthoniobacteraceae</i>
<i>Pedosphaeraceae</i>	0.51%	0.42%	<i>Rhodobacteraceae</i>	<i>Rhodospirillaceae</i>	0.31%	0.25%	<i>Saprospiraceae</i>	<i>Halothiobacillaceae</i>	0.47%	0.47%	<i>Acetobacteraceae</i>
<i>Gemmatimonadaceae</i>	0.46%	0.41%	<i>Peptococcaceae</i>	<i>Conexibacteraceae</i>	0.29%	0.25%	<i>Nostocaceae</i>	<i>Bifidobacteriaceae</i>	0.41%	0.45%	<i>Rhizobiaceae</i>

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