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**New nanoadsorbent based on magnetic iron oxide containing  
1,4,7,10-tetraazacyclododecane in outer chain ( $\text{Fe}_3\text{O}_4@\text{SiO}_2\text{-cyclen}$ ) for  
adsorption and removal of selected heavy metal ions  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$ ,  $\text{Cu}^{2+}$**

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## 18 **Abstract**

19 Magnetic Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-cyclen nanoparticles were prepared and used as adsorbent for Cd<sup>2+</sup>,  
20 Pb<sup>2+</sup> and Cu<sup>2+</sup> from aqueous solution removal process controlled with differential pulse anodic  
21 stripping voltammetry (DPASV) and hanging mercury drop electrode (HDME). Nanomaterial  
22 was synthesised in three-step process co-precipitation of Fe<sub>3</sub>O<sub>4</sub> core, coating with silane and N-  
23 (3-(triethoxysilyl)propyl)-1,4,7,10-tetraazacyclododecane-1-carboxamide silane  
24 functionalisation. The effectiveness of each step of the synthesis was confirmed using scanning  
25 electron microscopy (SEM), high-resolution X-ray photoelectron spectroscopy (XPS), powder  
26 X-ray diffraction (pXRD) and fourier-transform infrared spectroscopy (FT-IR) techniques. The  
27 Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-cyclen nanoparticles were employed for Cd<sup>2+</sup>, Pb<sup>2+</sup> and Cu<sup>2+</sup> ions elimination  
28 from individual and mixed solutions by carrying out titration with a suspension of  
29 nanocomposites. The binding level for all ions both in the individual solutions and in the  
30 mixture was very similar at high levels. For Cd<sup>2+</sup> and Cu<sup>2+</sup> ions sorption efficiency level was  
31 from 83% to 89%, while for Pb<sup>2+</sup> ions it was slightly lower at the level over 73%. In all cases,  
32 the equilibrium adsorption capacity parameter was over 1 mg/g and reached definitely higher  
33 values for individual ions solutions. The research results revealed that Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>- cyclen  
34 nanoparticles can be a promising adsorbent for magnetic heavy metal ions water treatment  
35 agents.

36

37 **Keywords:** functionalised Fe<sub>3</sub>O<sub>4</sub> nanoparticles, stripping voltammetry, Cd<sup>2+</sup>, Pb<sup>2+</sup> and Cu<sup>2+</sup>  
38 ions binding, nanoadsorbent, 1,4,7,10-tetraazacyclododecane, heavy metal ions removal

## 39 **Highlights**

- 40 • Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>- cyclen nanoparticles was utilised for the removal of Cd<sup>2+</sup>, Pb<sup>2+</sup>, Cu<sup>2+</sup>.
- 41 • The SEM, XPS, pXRD and FT-IR characterisation of nanoparticles was performed.
- 42 • Ions binding kinetics and efficiency were investigated with DPASV and HDME.
- 43 • Sorption efficiency was calculated and reached a maximum level of nearly 90%.
- 44 • The effectiveness of ions magnetic separation from the aqueous solution was confirmed.

45



## 46 1. Introduction

47 Over the last few decades, a significant increase has been observed in environmental  
48 pollution with heavy metal ions, which is a problem all over the world [1]. The increasing heavy  
49 metal ions concentration is generated mainly by technological and operational activities,  
50 production processes such as refining, power plants and combined heat and power plants, steel  
51 mills, coal and waste combustion, the chemical industry, the use of fertilisers and pesticides,  
52 and transport [2]. The presence of heavy metal ions in the environment constitutes a threat due  
53 to the possibility of their migration along the rungs of the trophic chain from soil, air, surface  
54 waters to plants, animals and in consequence to humans. Exposure to continuously increasing  
55 heavy metal ions concentrations lowers the action of biological barriers, which has a negative  
56 impact on human health due to their low biodegradability, bioaccumulation tendency, potential  
57 carcinogenic effect and mutagenicity [3]. Due to the danger posed by heavy metal ions  
58 environment contamination, the recycling of them from aquatic environments has become a  
59 very important area of scientific research and industry.

60 Until now, many methods of metal ions remediation from the environment have been  
61 used [4,5], particularly methods such as membrane filtration [6], osmosis [7], ion exchange [8],  
62 extraction with ionic liquids [9], photocatalysts [10], clays/layered double hydroxides [11],  
63 bioremediation [12], biomass and biosorption [13] phytoremediation [14], precipitation and  
64 coagulation [15,16] or electrocoagulation [17]. The conventional wastewater treatment  
65 technologies involving physical, chemical, and biological methods have their advantages but  
66 also disadvantages such as high costs, complex working operation, secondary pollutants  
67 production, low selectivity and many others [18]. To overcome these challenges, researchers  
68 used adsorption process as an alternative due to its wide working potentials, low price,  
69 simplicity in use, process efficiency, selectivity, capacity for material regeneration and  
70 pollutants trace concentration removal [19,20]. A wide range of materials have found  
71 application as heavy metal ions adsorbents e.g., chitosan and its functionalised forms [21],  
72 aerogels and their derivatives [22], based on cellulose [23], metal organic frameworks [24],  
73 carbon-based materials [25], polymers [26], dendrimers [27] and many others [28–31].  
74 However, among many research groups, various nanoscale adsorbents became very popular  
75 due to their high surface area, high mobility in solution, reactivity, good adsorption capacity  
76 and in consequence efficiency [32]. A series of nanoparticles and their modification were  
77 investigated as heavy metal adsorbents but, in recent years, a particular interest was attached to  
78 nanoparticles with magnetic properties, especially magnetite  $\text{Fe}_3\text{O}_4$  [33,34]. Magnetite

79 nanoparticles occupy a special importance due to the possibility of easy and quick isolation  
80 with the use of an external magnetic field [35]. Furthermore, creating core-shell type structures  
81 with silica  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  provides a good basis for further functionalisation thus for obtaining a  
82 material with wide applicability and selectivity [36–38]. There are known plenty of  
83  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  modified structures for heavy metal ions adsorption, however, the search for new  
84 and selective composites is still topical [39–41]. So far, there is lack of knowledge about crown  
85 ethers usage as surface nanocomposites modifiers. Qin et al. designed and synthesised silica gel  
86 bound di(aminobenzene)-18-crown-6(SGN18) for zirconium and hafnium adsorption [42]. Aza-  
87 crown ethers and derivatives in their heyday were widely used as complexation agents of metal  
88 ions in wide range of applications [43,44]. One of the well-known and widely used agent as a  
89 metal ions complexation was 1,4,7,10-tetraazacyclododecane [45–49].

90 In view of the fact that previously the authors used  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  nanoparticles  
91 containing nitrogen atoms in the outer chain as heavy metal ion adsorbent [50,51], they now  
92 decided to combine the excellent complexing properties of 1,4,7,10-tetraazacyclododecane and  
93 the possibility of pollutants magnetic isolation. In this work, magnetic core-shell nanoparticles  
94 functionalised with 1,4,7,10-tetraazacyclododecane  $\text{Fe}_3\text{O}_4@\text{SiO}_2$ -cyclen were designed,  
95 synthesised and used as heavy metal  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$ , and  $\text{Cd}^{2+}$  ions adsorbent for the first time. The  
96 newly obtained nanomaterial was characterised at every stage of synthesis using SEM, FT-IR,  
97 pXRD and (XPS) methods. Furthermore, the ions binding kinetics studies was performed and  
98 the binding capacity of nanoparticles was investigated using the differential pulse anodic  
99 stripping voltammetry (DPASV) method coupled with hanging mercury drop electrode (HDM)  
100 electrode. It is worth mentioning that the application of electrochemical detection skips  
101 conventional techniques' limitations in heavy metal detection and offers the same sensitivity  
102 with a lower cost, less complex operational procedures and fast on-site detection [52].

103

## 104 **2. Experimental**

### 105 **2.1. Reagents**

106 All analytical grade reagents were purchased from the indicated suppliers and used  
107 without further purification. All inorganic salts – ferric chloride hexahydrate ( $\text{FeCl}_3 \cdot 6\text{H}_2\text{O}$ ),  
108 ferrous chloride tetrahydrate ( $\text{FeCl}_2 \cdot 4\text{H}_2\text{O}$ ), potassium chloride KCl (99.9%), cadmium nitrate  
109 tetrahydrate  $\text{Cd}(\text{NO}_3)_2 \cdot 4\text{H}_2\text{O}$  (99.9%), lead nitrate  $\text{Pb}(\text{NO}_3)_2$ , and copper nitrate trihydrate  
110  $\text{Cu}(\text{NO}_3)_2 \cdot 3\text{H}_2\text{O}$  (99.9%) – were purchased from POCh (Poland). The organic solvents,  
111 ammonia (25%), tetraethylorthosilicate (98%) (TEOS), triethoxy(3-isocyanatopropyl)silane

112 were purchased from Sigma-Aldrich (Poland). 1,4,7,10-tetraazacyclododecane was bought  
113 from Strem Chemicals. Aqueous solutions were prepared using ultra-pure deionised water.

## 114 2.2. Synthesis of core-shell Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-cyclen nanostructures

115 New functionalised nanocomposites were obtained in the three-steps synthesis process.  
116 In the first stage, a magnetic Fe<sub>3</sub>O<sub>4</sub> core was received by the co-precipitation method. The  
117 mixture of Fe<sup>2+</sup>(FeCl<sub>2</sub>·4H<sub>2</sub>O) and Fe<sup>3+</sup> (FeCl<sub>3</sub>·6H<sub>2</sub>O) in aqueous solution using a three-neck  
118 flask was stirred under nitrogen conditions. Then the 25% ammonia solution was dripped into  
119 the reaction until the pH reached 11. Then the reaction mixture was heated to 60 °C for 3 hours  
120 to complete precipitation of the magnetite. The obtained nanoparticles were separated by a  
121 magnet and washed with distilled water several times.

122 The second stage of the synthesis involved the nanomagnetite coating with silica  
123 forming Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>, which was used for further modification. 16 g of Fe<sub>3</sub>O<sub>4</sub> nanoparticles  
124 obtained in the previous stage was dispersed in a mixture of 240 mL EtOH and 60 ml H<sub>2</sub>O.  
125 Then 3 mL of TEOS was added into reaction and the reaction was heated overnight to 50 °C.  
126 After this time, the obtained nanoparticles were magnetically collected and washed with  
127 distilled water several times.

128 In next stage of synthesis it was necessary to synthesise a modifier of the obtained  
129 nanoparticles Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub> due to the fact that the silane derivatives containing 1,4,7,10-  
130 tetraazacyclododecane are not a commercially available.

131 The reaction of 1,4,7,10-tetraazacyclododecane with triethoxy(3-  
132 isocyanatopropyl)silane was performed according to the modified procedure described  
133 previously [53] to obtain N-(3-(triethoxysilyl)propyl)-1,4,7,10-tetraazacyclododecane-1-  
134 carboxamide.

135 To the solution of 1,4,7,10-tetraazacyclododecane (4.179 g, 24,256 mmol) dissolved in  
136 200 ml dichloromethane 6 mL of triethoxy(3-isocyanatopropyl)silane (24,256 mmol) was  
137 dropwise added and the reaction was carried out overnight at 40 °C under a reflux condenser.  
138 Then the solvent was evaporated under reduced pressure. The obtained colourless oil was dried  
139 in a vacuum desiccator over P<sub>2</sub>O<sub>5</sub> to obtain 2.159 g of colourless solid with a yield of 21.33%.

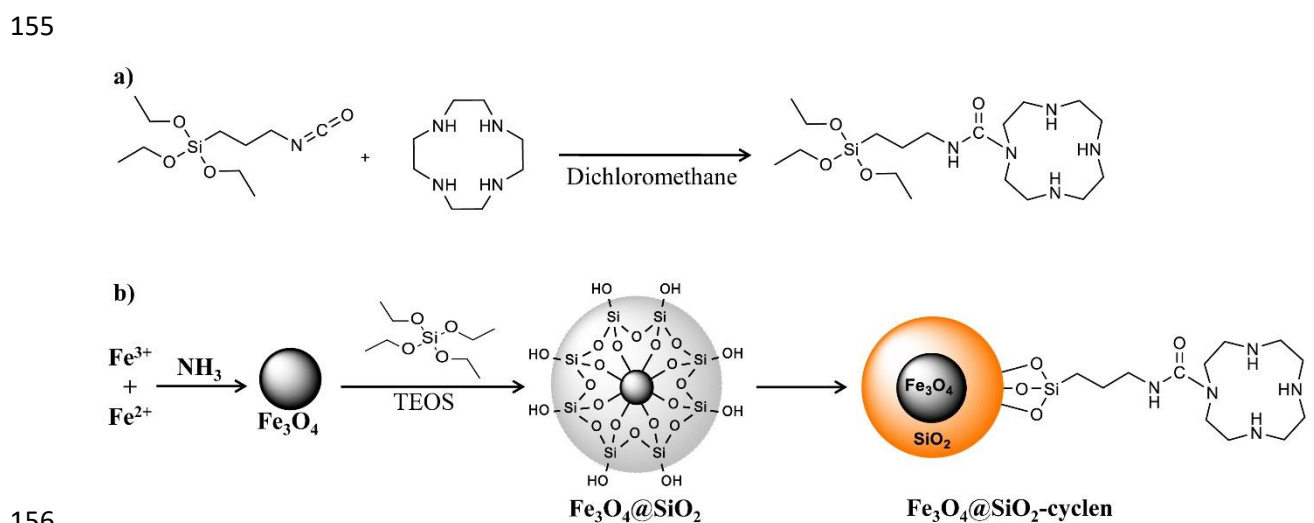
140 MALDI-TOF MS: m/z 419,29 [M+H]<sup>+</sup>, (MW = 420.31).

141 IR (KBr) (cm<sup>-1</sup>): 3291, 2972, 2926, 2886, 2735, 1627, 1536, 1478, 1444, 1391, 1365, 1254,  
142 1189, 1165, 1103, 1078, 955, 784.

143 <sup>1</sup>H-NMR (CDCl<sub>3</sub>): 0.538-0.572 (t, 2H, (CH<sub>3</sub>-CH<sub>2</sub>)<sub>3</sub>SiO-CH<sub>2</sub>-CH<sub>2</sub>-CH<sub>2</sub>-HN-, J<sub>1</sub> = 7 Hz;  
144 J<sub>1</sub> = 10,0 Hz; J<sub>2</sub>=8,5 Hz); 1.135-1.165 (dt, 9H, (CH<sub>3</sub>-CH<sub>2</sub>)<sub>3</sub>SiO-CH<sub>2</sub>-CH<sub>2</sub>-CH<sub>2</sub>-HN-, J<sub>1</sub>=1 Hz;

145  $J_1 = 1,5 \text{ Hz}$ ;  $J_1 = J_2 = 7.0 \text{ Hz}$ ); 1.479-1.623 (m, 2H,  $(\text{CH}_3\text{-CH}_2)_3\text{SiO-CH}_2\text{-CH}_2\text{-HN-}$ );  
 146 2.564-3.652 (m, 18H,  $(\text{CH}_3\text{-CH}_2)_3\text{SiO-CH}_2\text{-CH}_2\text{-CH}_2\text{-HN- -H-CH}_2\text{-CH}_2\text{-N-}$ ); 3.734-3.754 (q,  
 147 6H,  $(\text{CH}_3\text{-CH}_2)_3\text{SiO-CH}_2\text{-CH}_2\text{-CH}_2\text{-HN-}$ ,  $J_1=3 \text{ Hz}$ ;  $J_1 = 4 \text{ Hz}$ ;  $J_3 = J_4 = 3.0 \text{ Hz}$ );

148  
 149 The third stage consisted of  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  modification to obtain  $\text{Fe}_3\text{O}_4@\text{SiO}_2\text{-cyclen}$ .  
 150 1.593 g of  $\text{Fe}_3\text{O}_4@\text{SiO}_2$  was added to a solution of (3.324 g, 7,928 mmol) N-(3-  
 151 (triethoxysilyl)propyl)-1,4,7,10-tetraazacyclododecane-1-carboxamide dissolved in 150 mL  
 152 toluene. Then the reaction mixture was stirred for 12 h at 120 °C. After cooling to room  
 153 temperature the obtained product was magnetically collected and washed several times with  
 154 toluene and then with methanol. The resulting product was air-dried.



156  
 157  
 158 **Figure 1.** Scheme of a) nanoparticles modifier N-(3-(triethoxysilyl)propyl)-1,4,7,10-  
 159 tetraazacyclododecane-1-carboxamide and b) functionalised  $\text{Fe}_3\text{O}_4@\text{SiO}_2\text{-cyclen}$   
 160 nanocomposite synthesis.

161 **2.3. The  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$  and  $\text{Cu}^{2+}$  ions removal experiments from aqueous solutions by**  
 162  **$\text{Fe}_3\text{O}_4@\text{SiO}_2\text{-cyclen}$  binding**

163 The removal experiments of  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$  and  $\text{Cu}^{2+}$  ions from aqueous solutions were  
 164 performed by titration with  $\text{Fe}_3\text{O}_4@\text{SiO}_2\text{-cyclen}$  nanoparticles. The ions-binding abilities of  
 165  $\text{Fe}_3\text{O}_4@\text{SiO}_2\text{-cyclen}$  were studied in the solutions containing both individual ion or the mixture  
 166 of  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$  and  $\text{Cu}^{2+}$  ions in concentration level  $4.5 \cdot 10^{-6} \text{ M}$ . Before starting (DPASV)  
 167 measurement a 0.25 mg of  $\text{Fe}_3\text{O}_4@\text{SiO}_2\text{-cyclen}$  nanoparticles was added into ion solution at  
 168 each titration step and then mixed for 20 minutes All measurements were performed in 0.5 M  
 169 KCl pH 6.00.

## 170 2.4. Methods

171 Fourier Transform Infrared Spectroscopy (FT-IR) spectra were obtained were recorded  
172 on a Perkin Elmer Spectrum™ 3 FT-IR spectrometer with the KBr pellet technique.

173 Maldi-Tof mass spectra was performed on autoflex maX Bruker using  $\alpha$ -cyano-4-  
174 hydroxycinnamic acid (CCA) as a matrix.

175  $^1\text{H}$  NMR spectra were recorded in  $\text{CDCl}_3$  on a Bruker Advance III 500 MHz, where  
176 tetramethylsilane (TMS) was used as the internal standard. The chemical shifts are reported in  
177 ppm, while the signal patterns are given as follows: triplet (t), doublet of triplets (dt), quartet  
178 (q) and multiplet (m).

179 The scanning electron microscopy (SEM) images were captured with an FEI Quanta  
180 250 FEG (ThermoFisher Scientific) microscope, equipped with a Schottky field emission gun.  
181 The microscope was operating at 20 kV accelerating voltage in high vacuum mode.

182 High-resolution X-ray Photoelectron Spectroscopy (XPS) analysis was carried out using  
183 Escalab 250Xi multispectroscopy (ThermoFisher Scientific). The spectroscope is equipped  
184 with an  $\text{AlK}\alpha$  x-ray source; the pass energy through the hemispherical analyser was 20 eV. The  
185 measurement was assisted with low-energy electron and low-energy  $\text{Ar}^+$  ion bombardment for  
186 charge compensation. The measurements were carried out in C 1s, Fe 2p, Si 2p and N 1s core-  
187 level binding energy range, where the adventitious C 1s spectra (284.6 eV) was used for final  
188 peak calibration. Peak deconvolution was performed using Avantage v5.9921 (ThermoFisher  
189 Scientific).

190 The crystalline phases and crystallite size were determined by room-temperature  
191 powder x-ray diffraction (pXRD). The pXRD patterns were collected on D2 Phaser (Bruker)  
192 diffractometer, using  $\text{CuK}\alpha$  radiation ( $\lambda = 1.54056 \text{ \AA}$ ) with a scanning angle  $2\theta = 5\text{--}90^\circ$  and a  
193 step size  $0.01^\circ$ . LeBail refinement of the pXRD pattern was performed to determine the lattice  
194 parameters, using the DIFFRAC.SUITE TOPAS.

195 All differential pulse anodic stripping voltammetry (DPASV) measurements were  
196 performed using an Autolab potentiostat/galvanostat (PGSTAT-128N, Metrohm), equipped  
197 with NOVA 2.1.4 software. Three electrode Teflon electrochemical cell was used in all  
198 experiments to avoid metal ions sorption on the glass surface. A Static Drop Mercury Electrode  
199 (SDME) 663 VA Stand Metrohm was used as a working electrode. The reference electrode was  
200 calomel electrode  $\text{Hg}|\text{Hg}_2\text{Cl}_2|\text{KCl}_{(\text{saturated})}$  and platinum wire (Pt) were used as the counter  
201 electrode.



202 The detection of  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$ , and  $\text{Cd}^{2+}$  by DPASV method was performed using the  
 203 following conditions: deposition time 90 s, deposition potential  $-0.9$  V, modulation amplitude  
 204  $0.05$  V, modulation time  $0.07$  s, interval time  $1.85$  s, and step potential  $0.003$  V. The  
 205 simultaneous  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$ , and  $\text{Cd}^{2+}$  ions was performed in a potential range of  $-0.8$  V to  $-0.04$   
 206 V, while the individual detection of ions was performed in potential range of  $-0.8$  V to  $-0.5$  V,  
 207  $-0.6$  V to  $-0.3$  V and  $-0.4$  V to  $-0.04$  V for  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$ ,  $\text{Cu}^{2+}$  detection, respectively.

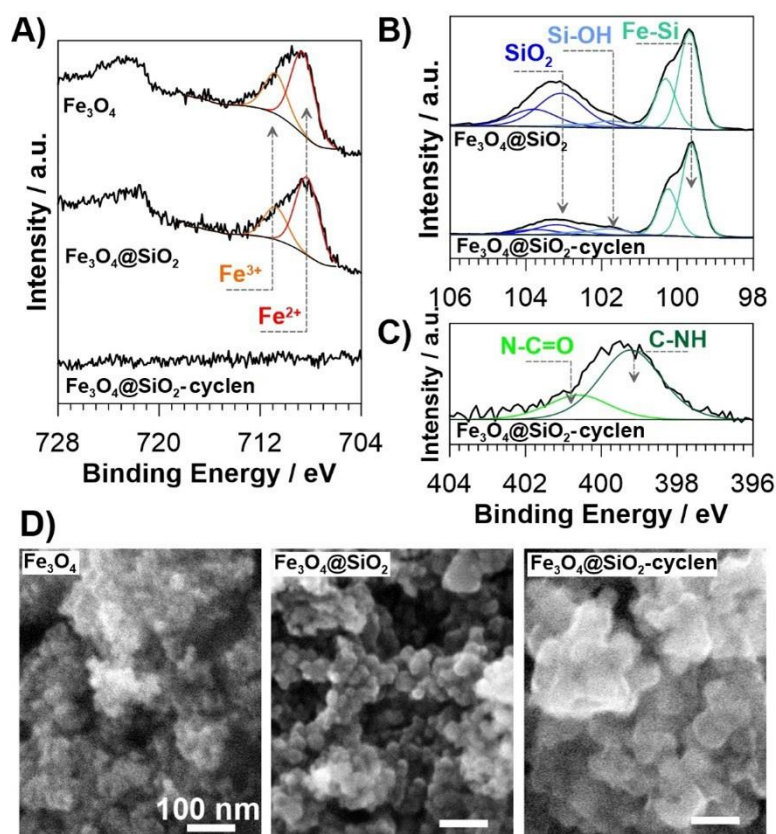
208 The metal ions solutions were prepared in the supporting electrolyte, which was  
 209 potassium chloride KCl  $0.5$  M, pH  $6.0$ . The  $\text{Fe}_3\text{O}_4@SiO_2$ -cyclen nanoparticles suspension in  
 210 supporting electrolyte solution before each measurement was freshly prepared by using  
 211 ultrasonic bath for  $30$  minutes.

212

### 213 3. Results and discussion

#### 214 3.1. XPS and SEM analysis

215 The high-resolution XPS analysis was performed to evaluate the efficiency of  
 216 consecutive  $\text{Fe}_3\text{O}_4$  nanoparticles functionalisation steps. The XPS spectra of each studied  
 217 sample are presented in Fig. 2.



218



219 **Figure 2.** High-resolution XPS analysis of the studied Fe<sub>3</sub>O<sub>4</sub>, Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub> and  
 220 Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-cyclen samples in the binding energy range of: A) Fe 2p, B) Si 2p and C) N 1s  
 221 peak with spectral deconvolution. D) SEM micrographs of studied materials. The scale bar  
 222 marks 100 nm.

223 Figure 2A shows iron chemistry. The spectra obtained for the Fe<sub>3</sub>O<sub>4</sub> sample may be  
 224 deconvoluted into two Fe 2p<sub>3/2</sub> features, peaking at 708.6 and 710.8 eV, which is characteristic  
 225 of Fe<sup>2+</sup> and Fe<sup>3+</sup> in Fe<sub>3</sub>O<sub>4</sub>, respectively [54]. The calculated Fe<sup>2+</sup>:Fe<sup>3+</sup> ratio is 2:1. Importantly,  
 226 upon the formation of Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub> core-shell structure, the peak position as well as the  
 227 Fe<sup>2+</sup>:Fe<sup>3+</sup> ratio remains unchanged, testifying to unchanged iron chemistry and structure within  
 228 the nanoparticles. Nevertheless, the Fe 2p spectra intensity weakens upon the NP coverage with  
 229 an SiO<sub>2</sub> shell, and eventually becomes indistinguishable for Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-cyclen, due to the  
 230 limited depth of the XPS signal origin (approx. 3-5 nm).

231 The Si 2p spectra shows more complex character, with not less than three deconvolution  
 232 peak doublets to be distinguished in Figure 2B. The characteristic component at 103.2 eV is  
 233 attributed to SiO<sub>2</sub> while the second, smaller component negatively shifted at 0.9 eV may  
 234 correspond to Si-OH bonds [55]. The third, notable component with Si 2p<sub>3/2</sub> at 99.8 eV may  
 235 have multiple origins, including Si-Fe bonds [56] or even Si-C bonds [57]. Adventitious carbon  
 236 contribution in Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub> NP powder is a relevant source for Si-C. After further modification  
 237 with N-(3-(triethoxysilyl)propyl)-1,4,7,10-tetraazacyclododecane-1-carboxamide the share of  
 238 99.8 eV feature rises by 35%. This modification is assisted with two time decreases of the SiO<sub>2</sub>  
 239 share, again due to the restricted depth of the XPS analysis. The detailed information are  
 240 presented in **Table 1**.

241  
 242 **Table 1.** The chemical composition (in at.%) using XPS analysis by peak deconvolution.

Sample	Fe 2p		Si 2p			N 1s	
	Fe <sup>2+</sup>	Fe <sup>3+</sup>	SiO <sub>2</sub>	Si-OH	Si-(Fe/C)	NH-C	NC=O
BE / eV	708.6	710.8	103.2	102.3	99.8	399.3	400.6
Fe <sub>3</sub> O <sub>4</sub>	67.1	32.9	--	--	--	--	--
Fe <sub>3</sub> O <sub>4</sub> @SiO <sub>2</sub>	1.3	0.6	40.2	6.9	50.0	--	--
Fe <sub>3</sub> O <sub>4</sub> @SiO <sub>2</sub> -cyclen	--	--	16.4	9.7	67.4	4.8	1.7

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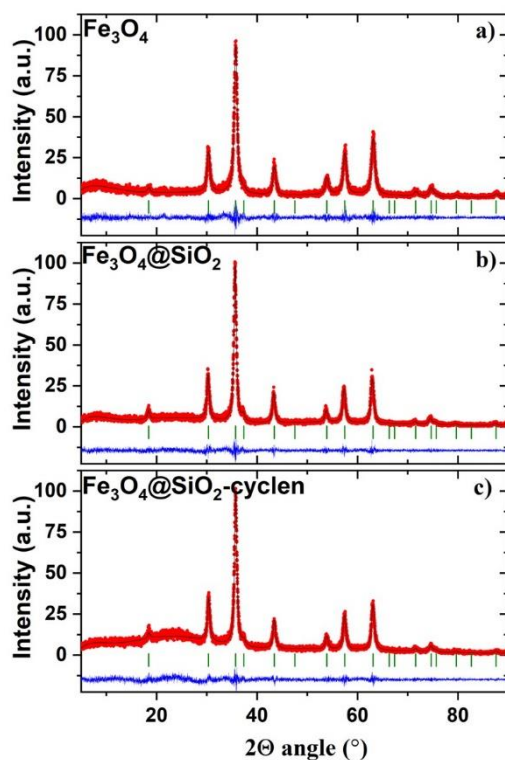
244 While the N-(3-(triethoxysilyl)propyl)-1,4,7,10-tetraazacyclododecane-1-carboxamide  
245 compounds do not introduce C 1s features that can be easily differentiated from the surface-  
246 adsorbed adventitious carbon under air exposure, the successful Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-cyclen synthesis  
247 may be confirmed tracking N 1s spectral shape. As seen in Figure 2C this peak was  
248 deconvoluted using two separate components. The first, and more notable, at 399.3 eV is  
249 characteristic of NH-C bonds [58], while the latter, at 400.6 eV, was assigned to NC=O [59].

250 The morphology of each studied nanomaterial was revealed by SEM micrographs as  
251 seen in Figure 2D. The size of the Fe<sub>3</sub>O<sub>4</sub> nanoparticles is below 20 nm, thus making it  
252 impossible for accurate estimation. Such small particle size is reported as sufficient to achieve  
253 superparamagnetism in Fe<sub>3</sub>O<sub>4</sub> NPs [60]. Further encapsulation in SiO<sub>2</sub> by TEOS allows for full  
254 NP contours to be recognised and increasing the average size to approx. 34 ± 5 nm. This result  
255 suggests the SiO<sub>2</sub> thickness to exceed 5 nm, testifying for nearly the entire removal of the high  
256 resolution Fe 2p signal. The final functionalisation notably enlarges the Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-cyclen  
257 dimensions, which now reach 50 to as much as 100 nm. The less regular size introduced by  
258 Sil- Prop- cyclen is the probable explanation behind a significant share of Si still to be  
259 recognised by XPS.

260

### 261 **3.2. pXRD analysis**

262 The pXRD analysis was used to investigate the crystallinity nature, and size of crystallites of  
263 bare Fe<sub>3</sub>O<sub>4</sub> nanoparticles and after each functionalisation steps. The XRD patterns of a) bare  
264 Fe<sub>3</sub>O<sub>4</sub>, b) silica-coated Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>, and c) functionalised Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-cyclen nanoparticles  
265 together with the LeBail refinement (solid red line) are presented in Figure 3. A difference  
266 plot (between experimental and fitted data) and the Bragg positions are also shown in Figure  
267 3.



268

269 **Figure 3.** Powder x-ray diffraction pattern (pXRD) (red points) together with the LeBail  
 270 refinement profile (solid black line) for a) bare  $\text{Fe}_3\text{O}_4$ , b) silica coated  $\text{Fe}_3\text{O}_4@SiO_2$  and c)  
 271 functionalized  $\text{Fe}_3\text{O}_4@SiO_2$ -cyclen. The green vertical bars indicate the expected Bragg peak  
 272 positions for  $\text{Fe}_3\text{O}_4$  (JCPDS card PDF 00-065-0731).

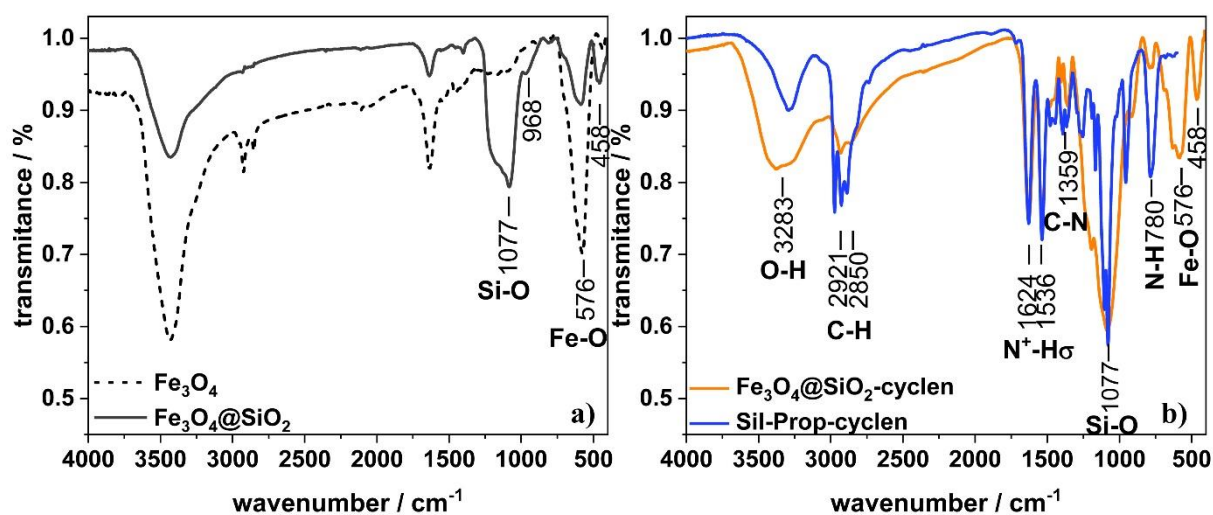
273

274 Our pXRD analysis confirms that for each studied sample all the diffraction peaks  
 275 correspond to the cubic crystal structure (space group  $Fd-3m$ , no. 227). Estimated lattice  
 276 parameters for the  $\text{Fe}_3\text{O}_4$  phase do not differ much and are  $a = 8.329(1) \text{ \AA}$  for bare  $\text{Fe}_3\text{O}_4$ ,  $a =$   
 277  $8.3521(6) \text{ \AA}$  for  $\text{Fe}_3\text{O}_4@SiO_2$ , and  $a = 8.3462(6) \text{ \AA}$  for  $\text{Fe}_3\text{O}_4@SiO_2$ -cyclen nanoparticles. The  
 278 average crystallite size of the  $\text{Fe}_3\text{O}_4$  phase was estimated from the Scherrer formula based on  
 279 the line broadening at half the maximum intensity of the reflection (220). The estimated  
 280 crystallite sizes are 14, 19 and 18 nm for  $\text{Fe}_3\text{O}_4$ ,  $\text{Fe}_3\text{O}_4@SiO_2$ , and  $\text{Fe}_3\text{O}_4@SiO_2$ -cyclen samples,  
 281 respectively. The obtained results are in agreement with SEM micrographs.

282

### 283 3.3. FT-IR spectroscopy analysis

284 FT-IR analysis was performed to identify the appropriate functional groups present in  
 285 the  $\text{Fe}_3\text{O}_4$  nanoparticles at each functionalisation steps.



286

287 **Figure 4.** FT-IR spectra for a) bare  $\text{Fe}_3\text{O}_4$ , silica coated  $\text{Fe}_3\text{O}_4@SiO_2$  and b) functionalised  
 288  $\text{Fe}_3\text{O}_4@SiO_2$ -cyclen nanoparticles, modifier silane N-(3-(triethoxysilyl)propyl)-1,4,7,10-  
 289 tetraazacyclododecane-1-carboxamide (Sil-Prop-cyclen).

290 The  $\text{Fe}_3\text{O}_4$  spectrum (Figure 4 a) shows strong stretching vibration Fe–O band at 576  
 291  $\text{cm}^{-1}$  typical for iron oxide [61]. Also observed was the reduction in the Fe–O band intensity for  
 292 coated nanoparticles with silica and functional group. The band intensity decrease confirms the  
 293 success in the surface functionalisation process [50]. After silica coating of nanomagnetite the  
 294 intense Si–O band appeared at  $1077\text{ cm}^{-1}$ . It is associated with stretching vibrations of Si–O–Si  
 295 and O–Si–O on the surface of the magnetic core at  $968\text{ cm}^{-1}$  and  $1077\text{ cm}^{-1}$ , respectively [62,63].  
 296 The broad strong band at  $3433\text{ cm}^{-1}$  is related to the overlapping of O–H stretching bond, which  
 297 proves hydroxyl groups formation on the magnetite surface [64]. Figure 4 b blue line shows  
 298 spectra for nanoparticles modifier N-(3-(triethoxysilyl)propyl)-1,4,7,10-  
 299 tetraazacyclododecane-1-carboxamide for comparison. A  $780\text{ cm}^{-1}$  weak band and a two headed  
 300 band about  $1536\text{ cm}^{-1}$  and  $1624\text{ cm}^{-1}$  characteristic for N–H vibrations are observed.  
 301 Furthermore, there are bands that appeared at  $1359\text{ cm}^{-1}$  and about  $2900\text{ cm}^{-1}$  corresponding to  
 302 C–N and N–H stretching vibrations, respectively. The same characteristic bands for the  
 303 spectrum for the functionalised nanoparticles  $\text{Fe}_3\text{O}_4@SiO_2$ -cyclen (Figure 4 b orange line) are  
 304 observed, which proves the success of the nanomaterials synthesis process. In addition, there is  
 305 a bandwidth extension at  $3283\text{ cm}^{-1}$  which probably corresponds to overlapping O–H and N–H  
 306 bands. On the basis of SEM, XPS and FT-IR analysis it can be concluded that the functionalised  
 307 core-shell nanoparticles formation was conducted successfully.

308

309

### 310 **3.3. Electrochemical analysis**

311 In this work, the independent and simultaneous determination of  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$  and  $\text{Cu}^{2+}$   
312 ions was performed under optimised experimental conditions using the DPASV method and  
313 HDM electrode. 1,4,7,10-tetraazacyclododecane functionalised magnetite nanoparticles  
314  $\text{Fe}_3\text{O}_4@\text{SiO}_2\text{-cyclen}$  were investigated as a new heavy metal ions adsorbent. To examine the  
315 ions-binding abilities of  $\text{Fe}_3\text{O}_4@\text{SiO}_2\text{-cyclen}$ , a series of measurements were performed in the  
316 solutions containing  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$  and  $\text{Cu}^{2+}$  ions in concentration level  $4.5 \cdot 10^{-6}$  M. All  
317 electrochemical experiments were performed in 0.5 M KCl pH 6.00. The pH was adjusted to  
318 avoid the formation of other metal ion species, mainly hydroxides, in the solution aside from  
319 the free ion [65–67]. The selection of pH and an influence of electrolyte was previously  
320 described by the authors in review paper [68]. There was three well-defined peaks observed at  
321 anodic stripping voltammograms at  $-0.65$  V,  $-0.44$  V and  $-0.20$  V confirming the presence of  
322  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$  and  $\text{Cu}^{2+}$  ion, respectively. Current intensities were multiplied by the dilution factor  
323 expressed by the equation (1):

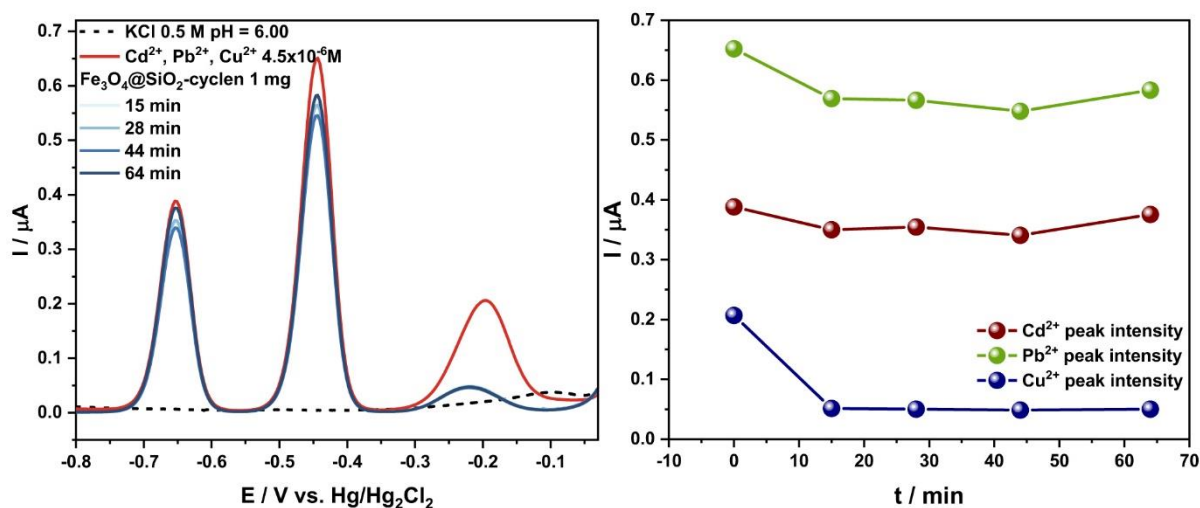
$$324 \quad DF = \frac{V_0 + V_s}{V_0} \quad (1)$$

325 where  $V_0$  is initial volume and  $V_s$  is the volume of added nanoparticles suspension in a  
326 particular portion.

#### 327 **3.3.1. $\text{Cd}^{2+}$ , $\text{Pb}^{2+}$ , $\text{Cu}^{2+}$ ions binding by $\text{Fe}_3\text{O}_4@\text{SiO}_2\text{-cyclen}$ - kinetics examination**

328 First, kinetic studies were performed to determine the effectual time to establish the  
329 equilibrium in the solution after adding the nanoparticles portion. All voltammograms were  
330 recorded immediately after adding 1 mg of  $\text{Fe}_3\text{O}_4@\text{SiO}_2\text{-cyclen}$  nanoparticles to the ion  
331 solution and after every 5 minutes of having mixed the solution. In the graph is shown the actual  
332 time that the nanoparticles stayed in the solution due to the length of the DPASV measurement  
333 (Figure 5).





334

335 **Figure 5.** Simultaneous differential pulse anodic stripping voltammograms for examination of  
 336 kinetics Cd<sup>2+</sup>, Pb<sup>2+</sup> and Cu<sup>2+</sup> (4.5·10<sup>-6</sup>M) ions binding – changes of peak intensities over time  
 337 after addition of 1 mg Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-cyclen nanoparticles.

338 After adding one portion of nanoparticles to the ion solution, a decrease in the peak  
 339 intensity in the case of only the Cu<sup>2+</sup> ion was observed. It can be stated that the addition 1 mg  
 340 of nanoparticles does not affect Cd<sup>2+</sup> and Pb<sup>2+</sup> ions binding. It can be assumed that the slight  
 341 peaks intensity changes are within the limit of the measurement error. On the basis of the  
 342 experiment it can be concluded that the equilibrium in the solution is established no more than  
 343 15 minutes after adding the portion of nanoparticles. A different phenomenon is observed in  
 344 the case of an experiment in independent Cd<sup>2+</sup>, Pb<sup>2+</sup> and Cu<sup>2+</sup> ion solutions (Figure S1. in the  
 345 Supporting Information file). In such a case, one observes a decrease in the peaks intensity over  
 346 time for both the Cd<sup>2+</sup> and Cu<sup>2+</sup> ion after adding 0.5 mg Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>- cyclen nanoparticles.  
 347 Functionalised nanomagnetite show no tendency to Pb<sup>2+</sup> ions binding and it is certainly not  
 348 influenced by the time of their presence in the solution. It can be assumed that in the case of  
 349 Cd<sup>2+</sup> and Cu<sup>2+</sup> ions the equilibrium in the solution is established after about 50 minutes after  
 350 adding the nanoparticles. After analysing the kinetics of ion binding by Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>- cyclen  
 351 nanoparticles, both in individual ion solutions and simultaneous analysis, it was found that the  
 352 optimal time for adding the nanoparticles portion to the solution to the measurement is 20  
 353 minutes after mixing. Also, the conducted kinetics experiment suggests that the ion binding  
 354 level may be influenced by the amount of added nanoparticles. Consequently, the observed

355 dependencies lead to the study of the Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>- cyclen nanoparticles binding capacity by  
356 titration of the Cd<sup>2+</sup>, Pb<sup>2+</sup> and Cu<sup>2+</sup> ions solution.

357

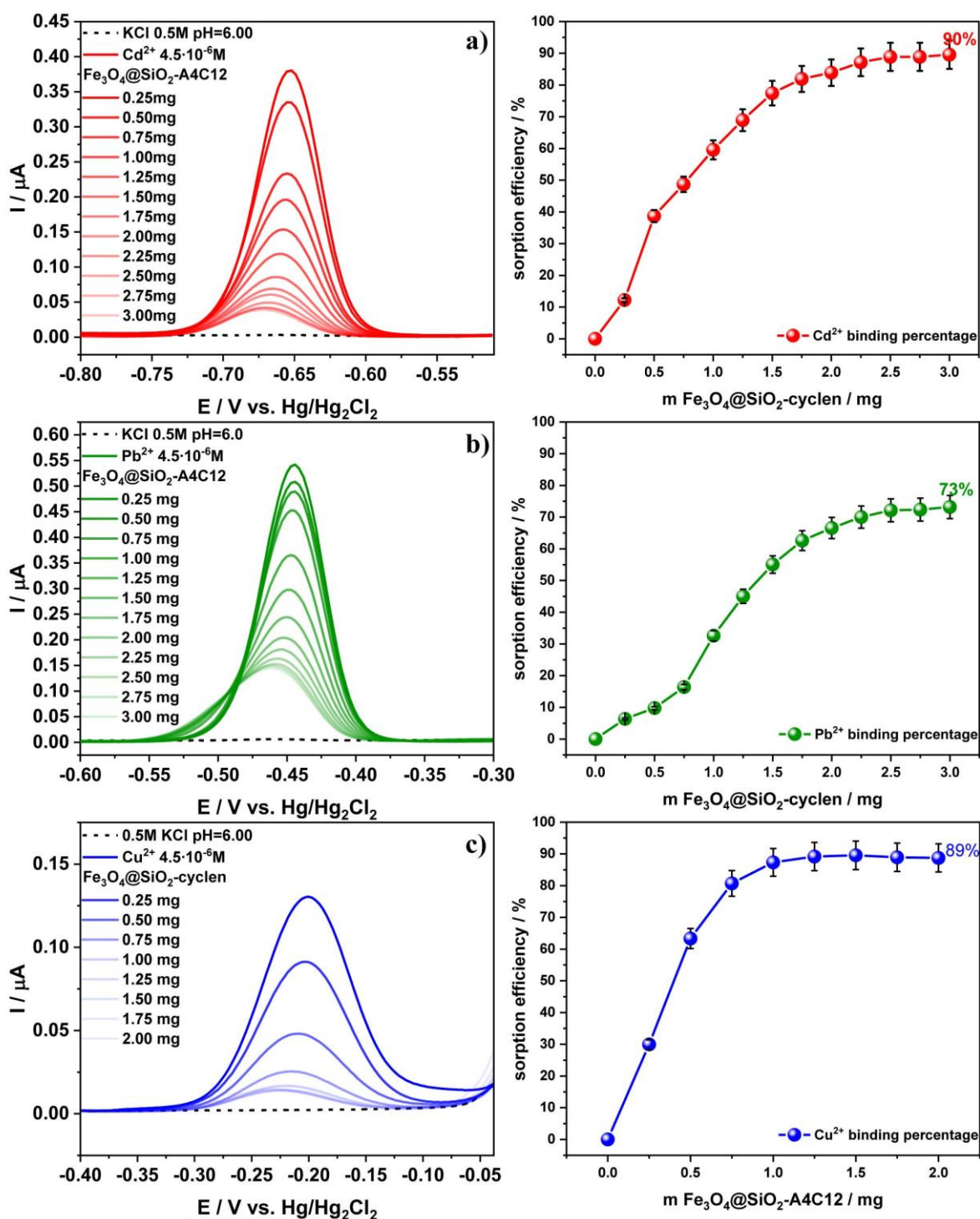
### 358 **3.3.2. Determination of Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-cyclen binding ability in Cd<sup>2+</sup>, Pb<sup>2+</sup> and Cu<sup>2+</sup> ions** 359 **solutions**

360 In order to determine the ion Cd<sup>2+</sup>, Pb<sup>2+</sup> and Cu<sup>2+</sup> binding capacity of  
361 Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>- cyclen nanoparticles, both in individual solutions and in the mixture, titration  
362 with nanoparticles was performed. Used as the measurement analytical method was a two-stage  
363 DPASV analysis involving pre-concentration and metal ions stripping. In the first stage, the  
364 negative potential of -0.9 V was applied to Cd<sup>2+</sup>, Pb<sup>2+</sup> and Cu<sup>2+</sup> ions electrodeposition on the  
365 working electrode. Subsequently, the faradic current obtained by oxidation was recorded during  
366 the potential sweep toward the positive potential direction from -0,8 V to -0.5 V, from -0,6 V  
367 to -0.3 V, from -0,4 V to -0.04 V and from -0.8 V to 0.0 V, for Cd<sup>2+</sup>, Pb<sup>2+</sup>, Cu<sup>2+</sup> and  
368 simultaneous ions detection, respectively. A 0.25 mg portion of Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>- cyclen  
369 nanoparticles was added into ion solution at each titration step and mixed for 20 minutes before  
370 starting the DPASV measurement. The mixing time was determined by the equilibrium  
371 establishment experiment in time. Peaks intensity changes during the addition of further  
372 nanoparticles portions were recalculated into sorption efficiency R<sub>%</sub> (%) at every measuring  
373 step based on the formula (2) [69]:

$$374 \quad R_{\%} = \frac{(C_0 - C_f)}{C_0} \cdot 100\% \quad (2)$$

375 where C<sub>0</sub> (mg/L) is initial ion Me<sup>2+</sup> concentration in the solution, C<sub>f</sub> (C<sub>e</sub>) (mg/L) is the final ion  
376 concentration remaining in the solution [69].





377

378 **Figure 6.** Individual differential pulse anodic stripping voltammograms and calculated sorption  
379 efficiency for a) Cd<sup>2+</sup>, b) Pb<sup>2+</sup> and c) Cu<sup>2+</sup> ions binding by Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-cyclen nanoparticles.

380

381 Figure 6 shows an individual Cd<sup>2+</sup>, Pb<sup>2+</sup> and Cu<sup>2+</sup> ions titration using  
382 Fe<sub>3</sub>O<sub>4</sub>@SiO<sub>2</sub>-cyclen nanoparticles in portion of 0.25 mg added in every titration step.

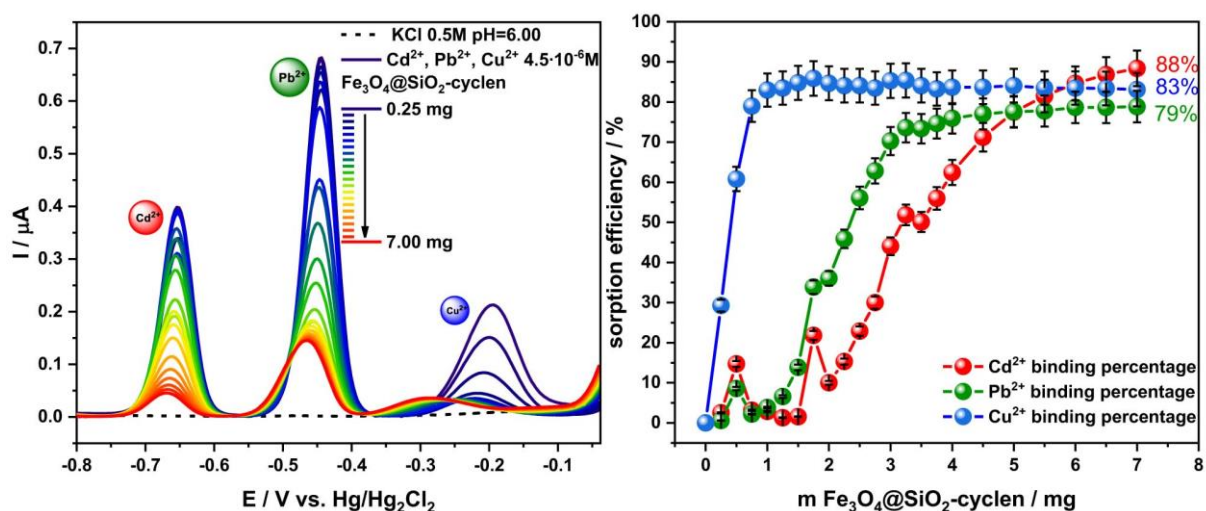
383 Figure 6 a) shows Cd<sup>2+</sup> ions binding capacity. While adding successive nanoparticles portions,





384 the peak intensity decreases, thus the sorption efficiency increases. After adding about 2–2.5 mg  
 385 of nanoparticles, the equilibrium is established and the next portions do not affect the peak  
 386 intensity reduction. The sorption effectiveness reaches about 90% and remains at this level  
 387 despite nanoparticles excess. A very similar phenomenon was observed during the titration of  
 388 the  $\text{Pb}^{2+}$  ion (Figure 6b). Initially, the peak intensity decreases with each subsequent  
 389  $\text{Fe}_3\text{O}_4@\text{SiO}_2$ -cyclen portion so that after the addition of about 2–2.5 mg of nanoparticles, it  
 390 reaches equilibrium. The titration ending determines the  $\text{Pb}^{2+}$  ion sorption efficiency at a level  
 391 of 73%. The  $\text{Fe}_3\text{O}_4@\text{SiO}_2$ -cyclen nanoparticles behaviour was very different in the  $\text{Cu}^{2+}$   
 392 solution. Similar as in the case of  $\text{Cd}^{2+}$  and  $\text{Pb}^{2+}$  ions titrations, a decrease in peak intensity  
 393 during titration was observed. However, a much smaller amount of nanoparticles was needed  
 394 to establish equilibrium in the solution. An equilibrium establishing was observed after the  
 395 addition of about 1 mg of nanoparticles. The addition of excess nanoparticles did not increase  
 396 the sorption efficiency, and its level reached about 89%.

397 Afterwards, to compare, all three ions were titrated simultaneously by  
 398  $\text{Fe}_3\text{O}_4@\text{SiO}_2$ -cyclen nanoparticles starting with 0.25 mg for the first portion and 0.5 mg in the  
 399 final steps (Figure 7).



400  
 401 **Figure 7.** Simultaneous differential pulse anodic stripping voltammograms and calculated  
 402 sorption efficiency for  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$ ,  $\text{Cu}^{2+}$  ions binding by  $\text{Fe}_3\text{O}_4@\text{SiO}_2$ -cyclen nanoparticles.

403 In the initial titration steps, only a decrease in the intensity of the  $\text{Cu}^{2+}$  peak was  
 404 observed. After adding about 1–1.5 mg of nanoparticles, the percentage of  $\text{Cu}^{2+}$  binding  
 405 remained unchanged at a level of 83%. Then, after the  $\text{Cu}^{2+}$  binding equilibrium establishing,  
 406 changes in the intensity of the  $\text{Cd}^{2+}$  and  $\text{Pb}^{2+}$  peaks were observed. As more  
 407  $\text{Fe}_3\text{O}_4@\text{SiO}_2$ -cyclen portions are added, the intensity of the  $\text{Cd}^{2+}$  and  $\text{Pb}^{2+}$  peaks was reduced

408 and the binding efficiency increased. The  $\text{Pb}^{2+}$  binding equilibrium was established after adding  
409 about 3–3.5 mg of nanoparticles to the ions solution and the sorption efficiency reached 79%.  
410 As the last equilibrium was  $\text{Cd}^{2+}$  binding established after the addition stage of ion binding  
411 remains at a comparable level, however, it requires different amount of nanoparticles.

412 The exact calculated sorption efficiency parameters for individual (from Figure 6) and  
413 simultaneous (from Figure 7) is presented in Table 2. The sorption efficiency  $R_{\%}$  calculated on  
414 the basis of the conducted DPASV experiments show that for  $\text{Cd}^{2+}$  this ion was detected at  
415 similar levels of 89-88% regardless as an individual ion in solution or in a mixture of studied  
416 ions. In the case of  $\text{Pb}^{2+}$ , the sorption level was higher for the simultaneous solution, while for  
417  $\text{Cu}^{2+}$  sorption value was higher in the single-component. Nevertheless, the sorption efficiency  
418 are at very high levels for all analysed solutions, which confirms high efficiency of ions binding  
419 by  $\text{Fe}_3\text{O}_4@\text{SiO}_2$ -cyclen nanoparticles.

420

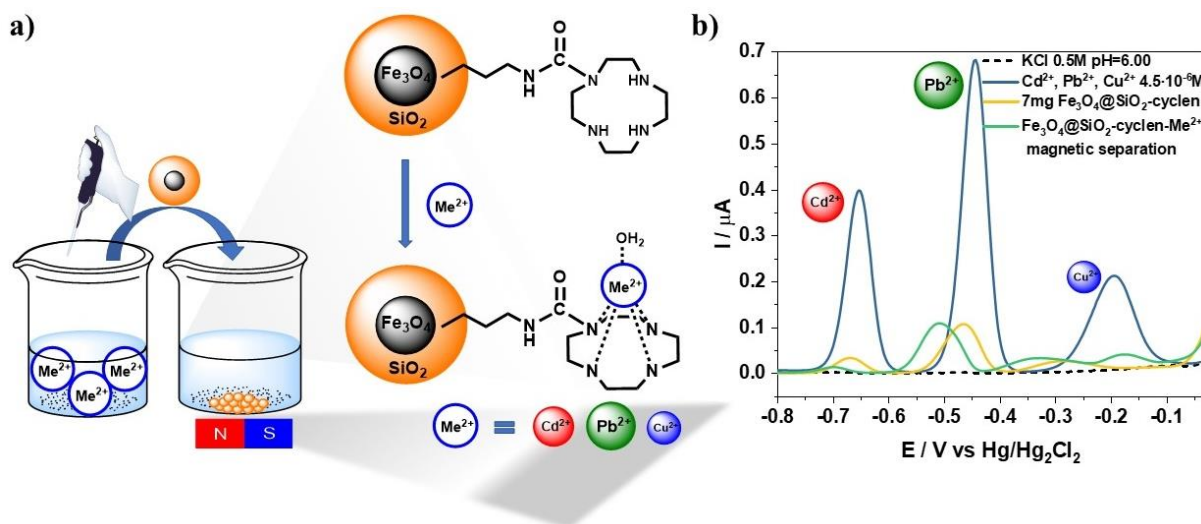
421 **Table 2.** Sorption efficiency parameters for individual and simultaneous  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$  and  $\text{Cu}^{2+}$   
422 ions binding by  $\text{Fe}_3\text{O}_4@\text{SiO}_2$ - cyclen nanoparticles.

	Individual	Simultaneous
	$R_{\%}$	$R_{\%}$
$\text{Cd}^{2+}$	89.55%	88.37%
$\text{Pb}^{2+}$	73.20%	78.85%
$\text{Cu}^{2+}$	88.75%	83.08%

423

### 424 3.3.3. Magnetic separation of adsorbed $\text{Cd}^{2+}$ , $\text{Pb}^{2+}$ and $\text{Cu}^{2+}$

425 Finally, the separation effectiveness of nanoparticles with adsorbed heavy metal ions by  
426 external magnetic field was investigated. First, the necessary amount of nanoparticles to bind  
427  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$ ,  $\text{Cu}^{2+}$  ions solution was added to them. After that, the reduction in the metal ions  
428 peaks intensity at the DPAS voltammograms was observed (from blue line to yellow line). Then  
429 the nanoparticles contained in the solution were isolated with the use of a magnet and the  
430 solution was separated and subjected to DPASV measurement. The obtained voltammogram,  
431 apart from slightly negatively shifted potentials, shows no deviation from that obtained directly  
432 after the adsorption process. This phenomenon indicates the separation of all bound ions along  
433 with the nanoparticles from the solution (Figure 8).



434

435 **Figure 8.** a) Scheme of magnetic  $\text{Me}^{2+}$  ions binding and magnetic separation experiment. Ion  
 436 binding model by aza-crown based on Subat et al. [46]. b) Simultaneous differential pulse  
 437 anodic stripping voltammograms: of  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$ ,  $\text{Cu}^{2+}$  ions (blue line), after adsorption by 7 mg  
 438  $\text{Fe}_3\text{O}_4@\text{SiO}_2\text{-cyclen}$  nanoparticles (yellow line) and after magnetic isolation nanoparticles with  
 439 adsorbed ions (green line).

440

#### 441 4. Conclusions

442 In this work, new core-shell nanostructures were synthesised containing a magnetic  
 443  $\text{Fe}_3\text{O}_4$  core and silane shell functionalised with 1,4,7,10-tetraazacyclododecane were  
 444 synthesised and characterised and their binding capacity was investigated. The results of  
 445 nanomaterial characteristics using SEM, XPS, pXRD and FT-IR confirmed the success of the  
 446 three-stage synthesis process.

447 Moreover,  $\text{Fe}_3\text{O}_4@\text{SiO}_2\text{-cyclen}$  nanoparticles binding kinetics and the sorption  
 448 efficiency of  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$  and  $\text{Cu}^{2+}$  was investigated both in individual ions solutions and in an  
 449 ions mixture. Electrochemical studies proved that nanoparticles can be effectively used as  
 450 adsorbents in environment heavy metal ions pollution management, which can be easily  
 451 collected using an external magnetic field.

452 The research showed that nanoparticles have a binding tendency to all selected ions  
 453  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$  and  $\text{Cu}^{2+}$  in their individual solutions and their sorption efficiency remains at 90%,  
 454 73% and 89%, respectively. It is worth mentioning that the smallest amount of nanoparticles  
 455 was needed to bind the  $\text{Cu}^{2+}$ . Simultaneous  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$  and  $\text{Cu}^{2+}$  determination also indicates  
 456 the ability of  $\text{Fe}_3\text{O}_4@\text{SiO}_2\text{-cyclen}$  to bind of all ions. However, a clear order of ion binding was

457 observed in the mixture during the addition of further nanoparticles portions. First,  $\text{Cu}^{2+}$  ions  
458 were bound, then  $\text{Pb}^{2+}$  and  $\text{Cd}^{2+}$  the sorption efficiency remains at 83%, 79% and 88% level,  
459 respectively. This phenomenon can be explained by the complexation constants value of these  
460 ions by 1,4,7,10-tetraazacyclododecane. The fastest  $\text{Cu}^{2+}$  binding can be explained by the  
461 highest value of the complexation constant 23.9, 24.8 according to Antunes et al. [70], whereas  
462 the complexation constants of  $\text{Pb}^{2+}$  and  $\text{Cd}^{2+}$  are 15.9 and 14.3, respectively [70]. Additionally,  
463 the level of ion binding depends on the amount of added nanoparticles. On this basis, it can be  
464 concluded that the presence of cyclen functional groups on the surface of nanoparticles gives  
465 them the ability to bind ions. Furthermore, the  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$ ,  $\text{Cu}^{2+}$  ions adsorption and magnetic  
466 separation experiment confirms the effectiveness of ion binding and purification of water  
467 environments. On the basis of this work, it can be assumed that the new nanomaterials based  
468 on  $\text{Fe}_3\text{O}_4$  and functionalised with N-(3-(triethoxysilyl)propyl)-1,4,7,10-tetraazacyclododecane-  
469 1-carboxamide can be successfully used as a universal  $\text{Cd}^{2+}$ ,  $\text{Pb}^{2+}$ ,  $\text{Cu}^{2+}$  ions magnetic  
470 adsorbent.

471

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## 475 **CRedit authorship contribution statement**

476 **Amanda Kulpa-Koterwa:** Methodology, Formal analysis, Investigation, Data Curation,  
477 Writing - Original Draft, Writing - Review & Editing, Visualization, Funding acquisition **Jacek**  
478 **Ryl:** Investigation, Data Curation, Writing - Original Draft, **Karolina Górnicka:** Investigation,  
479 Data Curation, Writing - Original Draft **Paweł Niedziałkowski:** Conceptualization, Resources,  
480 Writing - Original Draft, Writing - Review & Editing, Supervision

## 481 **Declaration of competing interest**

482 The authors declare that they have no known competing financial interests or personal  
483 relationships that could have appeared to influence the work reported in this paper.

484



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